

Appendix C

Computation of Detection and Quantitation Limits

INTRODUCTION

This appendix supports the Technical Support Document (TSD) for EPA's peer review of detection and quantitation concepts. It presumes that the reader has read Chapters 3 - 5 of the TSD.

This appendix compares detection and quantitation limits computed from data gathered by EPA. The comparison shows that, in general, detection limits derived from a single concentration such as EPA's MDL are, on average, approximately the same as detection limits derived from similar concepts such as the ACS LOD and LOQ and ISO/IUPAC CRV and MDV, and are approximately three times lower than a single-laboratory variant of ASTM's IDE; and that all quantitation limit concepts, such as EPA's ML, the ACS and ISO/IUPAC LOQ, and a single-laboratory variant of ASTM's IQE, produce approximately the same quantitation limits.

BACKGROUND

The regulated industry has been commenting on EPA's method detection limit (MDL; see the definition at 40 CFR 136, appendix B) since 1985 and on EPA's minimum level of quantitation (ML; see, e.g., the definition in the glossary at the end of EPA Method 1631C promulgated at 40 CFR 136, appendix A) since about 1987. In the early 1990s members of the regulated industry community began providing suggestions for alternate detection and quantitation limit concepts, most notably the compliance monitoring detection level (CMDL), compliance monitoring quantitation level (CMQL) and the alternate minimum level (AML). Most recently the industry community has advanced concepts for an interlaboratory detection estimate (IDE) and interlaboratory quantitation estimate (IQE) through ASTM-International (ASTM).

In response to industry's comments, EPA began a data gathering activity designed to allow evaluation of the various detection/quantitation limit concepts. The data would be used to characterize measurement variability versus concentration for the analytes most commonly measured in, and the analytical technologies most commonly employed for, environmental measurements.

SETTLEMENT AGREEMENT

In October of 2000, EPA signed an agreement associated with promulgation of EPA Method 1631 to settle a lawsuit brought by certain members of the regulated industry (the "Settlement Agreement"). The Settlement Agreement required EPA to, among other things, reassess detection and quantitation limit concepts, perform a peer review of the reassessment, and if warranted, propose changes to the MDL and ML or propose alternate concepts. This document supports EPA's reassessment.

EPA'S APPROACH TO ESTABLISHING DETECTION AND QUANTITATION LIMITS IN ANALYTICAL METHODS

The Engineering and Analysis Division (EAD) within EPA's Office of Science and Technology develops analytical methods for use in EPA's Clean Water Act (CWA) programs. In developing these methods, EAD first conducts a single-laboratory study in which an MDL and ML are determined followed by multiple single-laboratory studies in which the MDL and ML are either verified or if necessary, revised. Then, if resources, time, and applications of the method warrant, an interlaboratory study is conducted in which the MDL and ML are further verified or, if necessary, revised. To establish

that an MDL is realistic, EAD generally selects the highest MDL from among the MDLs determined by laboratories in the various studies or that can be verified by laboratories in the studies. For example, EPA determined the MDL in Method 1631 (mercury by cold-vapor atomic fluorescence) as 0.05 ng/L in a single laboratory and revised this MDL to 0.2 ng/L based on multiple single-laboratory studies. All laboratories verified the MDL of 0.2 ng/L in an interlaboratory study. Thus, although critics of the MDL have complained that the MDL is a single-laboratory concept that is highly variable and produces unrealistically low MDLs, the process used by EAD to establish the MDL and resulting ML is conservative and protects against unrealistically low MDLs and MLs.

EPA's VARIABILITY VERSUS CONCENTRATION STUDIES ("EPISODE 6000")

In 1997 and 1998, EPA conducted a study of variability vs. concentration for a number of analytical methods. Six laboratories were employed for the analyses; each analyte and method combination was tested by one of these laboratories. Details of the study design are described in EPA's *Study Plan for Characterizing Variability as a Function of Concentration for a Variety of Analytical Techniques* (July 1998). Based on the sampling episode number assigned to the study by the EPA Sample Control Center, the study and results have become known as the Episode 6000 study and data.

The analytes and analytical techniques studied were:

- Total suspended solids (TSS) by gravimetry,
- Metals by graphite furnace atomic absorption spectroscopy (GFAA),
- Metals by inductively-coupled plasma atomic emission spectrometry (ICP/AES),
- Hardness by ethylene diamine tetraacetic acid (EDTA) titration,
- Phosphorus by colorimetry,
- Ammonia by ion-selective electrode,
- Volatile organic compounds by purge-and-trap capillary column gas chromatography with a photoionization detector (GC/PID) and electrolytic conductivity detector (GC/ELCD) in series,
- Volatile organic compounds by gas chromatography with a mass spectrometer (GC/MS),
- Available cyanide by flow-injection/ligand exchange/amperometric detection,
- Metals by inductively-coupled plasma spectrometry with a mass spectrometer (ICP/MS),

In the study, an initial (range finding) MDL was determined for each combination of analyte and analytical technique using a revised draft of the MDL procedure. The revised draft had three significant changes: 1) the definition was more closely conformed to the MDL procedure; 2) optional iterative step 7 of the MDL procedure was made mandatory; and 3) the spike concentration to MDL ratio was reduced from 5 to 3 in an attempt to narrow the resulting MDL. During data gathering two laboratories complained that the reduction in spike to determined MDL ratio from 5 to 3 caused a large number of iterations and stated that 5 was more reasonable. Subsequently, EPA returned to the spike to MDL ratio of 5 published in the 40 CFR 136, Appendix B procedure.

After determining the initial MDL, each laboratory analyzed 7 replicates of samples spiked at concentrations of 100, 50, 20, 10, 7.5, 5.0, 3.5, 2.0, 1.5, 1.0, 0.75, 0.50, 0.35, 0.20, 0.15, and 0.10 times the initial MDL. In a few instances laboratories analyzed more than 7 replicates. Results associated with the replicate analyses at each concentration level were obtained, as often as possible, using the same calibration that was used in determining the initial MDL. Where laboratory reports indicated that multiple calibrations were conducted, the association between each result and its calibration was used in the data analysis.

Spiked aqueous solutions were analyzed in order from the highest concentration (100 times the MDL) to the concentration at which 3 or more non-detects (zeros) were encountered among the 7

replicates, or the lowest concentration specified (0.1 times the MDL), whichever occurred first. This analysis order (1) minimized carryover that could occur in some methods if a low-concentration sample had followed a high-concentration sample (as may happen when samples are analyzed in random order), and (2) prevented collection of a large number of zeros if the signal disappeared.

A variant of the iterative MDL procedure was used for organic compounds determined by chromatographic methods. Methods for organics normally list many (15 to 100) analytes, and the response for each analyte is different. Therefore, to determine an MDL for each analyte, the concentration of the spike would need to be inversely proportional to the response. Making a spiking solution with 15 to 100 different concentrations is cumbersome and error prone. The approach used in the study was to run 7 replicates at decreasing concentrations until signal extinction, then select the concentration(s) appropriate for the MDL for each analyte according to the MDL procedure. In some cases the laboratories selected the concentrations, in others cases, EPA did. This approach was generally applied for organics analysis. However, laboratories also had the option of using some combination of monotonically decreasing concentrations described above and a few selected concentrations to achieve the desired spiking levels.

DATA ANALYSIS

Data from the study were used to evaluate detection and quantitation limit concepts that employ a point estimate or employ a model of variability versus concentration. Concepts that were evaluated and that employ a point estimate were the:

- EPA method detection limit (MDL) and minimum level of quantitation (ML),
- American Chemical Society (ACS) limit of detection (LOD) and limit of quantitation (LOQ), and
- International Standards Organization/International Union of Pure and Applied Chemistry (ISO/IUPAC) critical value (CRV), detection limit (also termed "minimum detectable value" (MDV)), and LOQ.

Concepts that were evaluated and employ a model of variability versus concentration were the ASTM interlaboratory detection estimate (IDE) and interlaboratory quantitation estimate (IQE).

COMMONALITY OF CONCEPTS

The EPA, ACS, and ISO/IUPAC concepts are all multiples of the standard deviation of either replicate measurements of a blank or of the lowest spike concentration that produces positive (non-zero) results for all 7 replicates. Although some would argue that this difference is significant, in practice they are functionally analogous because a non-zero result is needed to compute any concept (a zero result will return zero as the detection or quantitation limit).

Other subtle distinctions are that (1) ISO/IUPAC suggest a false positive rate of 5 % ($\alpha = 0.05$) for the CRV and MDV, whereas EPA specifies a false positive rate of 1 % ($\alpha = 0.01$) for the MDL and (2) the EPA MDL was calculated by pooling data from two concentration levels after determining that the variabilities of the two concentration levels are not significantly different (as stipulated in step 7 of the revised MDL procedure), thereby increasing the degrees of freedom to 12 from the 6 used in computation of the ISO/IUPAC CRV and ACS LOD. The consequence of distinction 1) is that a concept with a higher allowed false positive rate ($\alpha = 0.05$) will produce a lower detection limit than a concept with a lower false positive rate ($\alpha = 0.01$). The consequence of distinction 2) is that a detection limit resulting from pooling at two levels will be lower and more stable than a detection limit at a single level (given the same variability at each level) because the degrees of freedom are increased in the t statistic.

The ACS and ISO/IUPAC concepts specify replicate measurements of a blank. In computing detection and quantitation limits from the Episode 6000 data, if the blank returned non-zero results, the concepts were computed using replicate measurements of the blank. If the blank returned a zero result in any of the 7 measurements, the lowest spike concentration (or, in the case of the MDL, two lowest spike concentrations) that produced a non-zero result was used for computation of all concepts. This simplification condensed the EPA MDL and the ACS LOD to a single concept subsequently termed the EPA/ACS DL. Similarly, the EPA ML and ACS LOQ were condensed to a single concept, termed the EPA/ACS QL.

The remaining single-point concepts were the ISO/IUPAC CRV, MDV, and LOQ. The ISO/IUPAC CRV differs from the EPA/ACS DL because of its suggested use of a false positive rate of 5% ($\alpha = 0.05$) versus use of a false positive rate of 1% ($\alpha = .01$) in the EPA/ACS DL. The ISO/IUPAC MDV also differs from the EPA/ACS DL because of (1) its suggested use of a false positive rate of 5% ($\alpha = 0.05$), (2) its false negative rate of 5% ($\beta = 0.05$), and (3) recovery correction (estimated using a linear regression). Therefore, the ISO/IUPAC CRV and MDV were each treated separately (were not combined with another concept) from the other detection limit concepts in the data analysis. The ISO/IUPAC LOQ is also different from the other quantitation limit concepts and was treated separately from these concepts.

The ASTM IDE and IQE were treated separately because they are constructed by fitting a model to variability versus concentration data, rather than being derived from the standard deviation of replicate measurements of a single concentration, as are the EPA, ACS, and ISO/IUPAC concepts. Similar to some of the ISO/IUPAC concepts, the ASTM IDE and IQE include protection against false negatives and recovery correction. The IQE, but not IDE, also includes an added correction for the bias associated with an estimate of the true standard deviation at each concentration. In the context of the IQE, the word "bias" means the amount by which the estimated sample standard deviation is low compared to the true population standard deviation, and should not be confused with common use of the word "bias" in an analytical measurement.

SINGLE-LABORATORY VARIANTS OF INTERLABORATORY CONCEPTS

EPA's Episode 6000 database contains single-laboratory data because of the prohibitive expense that would have been incurred in gathering interlaboratory data. Because the EPA, ACS, and ISO/IUPAC concepts are single-laboratory concepts, and the ASTM IDE and IQE are interlaboratory concepts, the ASTM concepts could not be computed using the single-laboratory data in the Episode 6000 studies. To solve this problem, single-laboratory variants of the IDE and IQE were used. These single-laboratory variants were termed the SL-IDE and SL-IQE for "single-laboratory IDE" and "single-laboratory IQE." The SL-IDEs and SL-IQEs were constructed using the overall standard deviation within a single laboratory at each concentration rather than the overall standard deviation across all laboratories at each concentration.

ATTEMPTED APPLICATION TO INTERLABORATORY DATA

EPA attempted to apply the various concepts to interlaboratory study data in response to a request by the Petitioners to the Settlement Agreement and so that detection and quantitation limits could be compared. However, because the EPA, ACS, and ISO/IUPAC concepts are single-laboratory concepts whereas the ASTM concepts are interlaboratory concepts, it was not possible to compute directly comparable detection and quantitation limits from the same data.

What was possible was to compare detection and quantitation limits produced by EPA and the Electric Power Research Institute (EPRI) for the EPA Method 1631 and EPA Method 1638

interlaboratory study data. Although the resulting detection and quantitation limits are single- or interlaboratory, as appropriate to the particular concept; i.e., apples versus oranges, their magnitudes are informative. The EPRI detection and quantitation limits are from EPRI reports of the results of the Method 1631 and Method 1638 studies.

COMPUTATIONS

All computations were carried out using Statistical Analysis System (SAS) version 8.01. The equations for all concepts were programmed into the SAS software by a senior statistician, with assistance from senior analysts. There is some ambiguity in the IUPAC/ISO and ASTM detection and quantitation limit concepts and in interpretation of results from the ASTM concepts. Several formulas are given in the IUPAC/ISO documentation, but none are defined to be the official ISO/IUPAC detection and quantitation limit concepts. Therefore, calculations for the CRV, MDV, and LOQ were chosen because they were most representative of Lloyd Currie's definitions of a critical value, detection limit and quantitation limit. The specific equations used are on the CD-ROM that supports this Appendix. Ambiguity in results from the ASTM concepts is attributable to the subjective nature of interpreting residual plots for each analyte. To resolve this issue, IDE and IQE models were chosen using significance tests for slope and curvature.

References used for the IUPAC/ISO concepts were those published by Currie in *Pure and Applied Chemistry* **67**:10, 1699-1723 (1995) as updated by *Analytica Chimica Acta* **391** 105-126 (1999). Where needed, the ASTM concepts were programmed as single-laboratory variants of the Practices D 6091 (IDE) and D 6512 (IQE). EPA has included the SAS program code on the CD-ROM that supports this document.

DATA SETS EVALUATED

EPA computed EPA/ACS detection limits and quantitation limits; ISO/IUPAC CRVs, MDVs and LOQs; and single-laboratory variants of ASTM IDEs (SL-IDEs) and IQEs (SL-IQEs) for the Episode 6000 data. EPA also computed IDEs and IQEs for the Method 1631 and 1638 interlaboratory study data.

DATA SETS NOT EVALUATED

The Petitioners and Intervenor to the Settlement Agreement provided the list of data sets shown in Table 1 and suggested that EPA evaluate detection/quantitation limit concepts using the data sets on the list. However, in reviewing the data sets suggested, EPA determined that many were developed for characterizing the behavior of an analyte or analytes across the analytical range of a method, rather than in the region of detection and quantitation, while others did not result from the IDE and IQE procedures. For example, any data set developed prior to the advent of the IDE and IQE would be inappropriate because there could not have been an estimate of IDE_0 or IQE_0 . This eliminates all data sets in Table 1 except the EPA/EPRI Method 1631 and Method 1638 data set, and the MMA 2001-2 data set. It is possible that some value in one or more of the data sets developed prior to the advent of the IDE and IQE would fortuitously meet the IDE/IQE criteria. But the IDE and IQE can be circular; i.e., once developed from a given data set there may be a value in the data set than can be construed to meet the criteria. The point is that data sets developed without following the IDE and IQE procedures, particularly without making an *a priori* estimate of IDE_0 or IQE_0 , do not meet the requirements of the IDE and IQE procedures, regardless of whether the data in them can be construed to have met those requirements after the fact.

In addition, these data sets do not lend themselves to the comparisons used in this report because 1) they contain interlaboratory data that cannot be reduced to single-laboratory data without treating each

laboratory separately¹ and 2) the developers of these data sets did not apply EPA's procedure and measurements for establishing an MDL and ML, so an MDL and ML could not be determined for making comparisons (see the section titled "EPA's Approach to Establishing Detection and Quantitation Limits in Analytical Methods").

Further, the EPA 6000 data set is comprehensive in coverage of analytes, analytical techniques, and a concentration range from 0.1 to 100 times the MDL, whereas the data sets suggested by Petitioners focus on metals, one organic analyte (PCBs), and concentrations across the analytical range of the method. The range of data used for construction of an IDE or IQE is particularly important. As detailed in the discussion of the "Effect of number and spacing of concentrations for determination of the SL-IDE and SL-IQE" below, including data across the analytical range in calculation of an SL-IDE significantly raises the SL-IDE.

As stated in the section titled "Attempted Application to Interlaboratory Data" EPA used the EPA Method 1631 and 1638 data sets in computations of detection and quantitation limits, with the qualifiers given in that section. The EPA Method 1631 and Method 1638 data sets were the only data sets suggested by the Petitioners that were used.

RESULTS OF COMPUTATIONS

Detection and quantitation limits are presented in a set of tables for the Episode 6000 study and a single table for the Method 1631 and Method 1638 studies. Within the Episode 6000 data set, results for detection limits are compared followed by results for quantitation limits. Within the comparison of limits (detection or quantitation), the first table compares the actual limits followed by a table of ratios between limits. These tables are followed by a rank comparison table, making a total of five tables for this data set.

EPISODE 6000 DATA

Table 2 compares detection limits produced by the four concepts (EPA/ACS DL; ISO/IUPAC CRV; ISO/IUPAC MDV; and ASTM SL-IDE) and Table 3 compares the ratio between these concepts, taking the EPA/ACS DL as reference. The ISO/IUPAC CRV was greater than the corresponding EPA/ACS DL for 26% of the analytes and methods. The median ratio of ISO/IUPAC CRV to EPA/ACS DL was significantly less than 1 based on the sign test with $\alpha = 0.05$ ($p < 0.0001$). We believe that the major reason for the difference is the different Type I error rate for the two concepts ($\alpha = 0.01$ for the EPA/ACS DL and $\alpha = 0.05$ for the ISO/IUPAC CRV).

The median ratio between the ISO/IUPAC MDV and the EPA/ACS DL is 1.2; i.e., the ISO/IUPAC MDV is a median of 1.4 times higher than the EPA and ACS concepts. The ISO/IUPAC MDV was greater than the corresponding EPA/ACS DL for 57% of the analytes and methods. The median ratio of ISO/IUPAC MDV to EPA/ACS DL did not differ significantly from 1 based on the sign test with $\alpha = 0.05$ ($p = 0.055$). The likely reason that the two concepts do not yield significantly different results is that the correction for false negatives and recovery correction in the MDV ($\beta = 0.05$) are counteracted by the smaller Type I error rate for the EPA/ACS DL.

¹ Treating each laboratory separately would lead to further ambiguities because results from some laboratories could produce detection and quantitation limits greater than the single-concentration limits whereas results from other laboratories could produce detection and quantitation limits less than the single-concentration limits. Given the variability of the data, such an outcome would be virtually assured and would provide no further useful information.

The median ratio between the ASTM SL-IDE and the EPA/ACS DL is 2.9; i.e., the single-laboratory variant of the ASTM IDE is a median of 2.9 times higher than the EPA and ACS concepts. The SL-IDE was greater than the corresponding EPA/ACS DL for 91% of the analytes and methods. The median ratio differed significantly from 1 based on the sign test with $\alpha = 0.05$ ($p < 0.0001$). The reason(s) for the difference could not be determined easily because of the number of confounding factors included in the ASTM SL-IDE.

Table 4 compares quantitation limits produced by the three concepts (EPA/ACS QL; ISO LOQ; and ASTM SL-IQE) and Table 5 compares the ratio between these concepts taking the EPA/ACS QL as reference. The median ratio between the ISO/IUPAC LOQ and the EPA/ACS QL is 0.92, and the median ratio between the ASTM SL-IQE and the EPA/ACS QL is 1. The ISO LOQ and ASTM SL-IQE are greater than the corresponding EPA/ACS QL for 43% and 51% of the analytes and methods, respectively. The median ratio did not differ significantly from 1 based on the sign test at $\alpha = 0.05$ (LOQ: $p = 0.062$; SL-IQE: $p = 0.78$). The reason(s) why the ASTM SL-IQE, ISO LOQ, and the EPA/ACS QL produce nearly identical limits could not be determined easily because of the number of confounding factors included in the ASTM SL-IQE.

Some of the differences or similarities in median quantitation limits may be accounted for by rounding in the ML procedure, although the rounding should average over the large number of analytes examined.

Table 6 gives the frequency comparisons for the detection and quantitation limits; i.e., the frequency with which each concept produced the highest or lowest quantitation limit.

EPA/EPRI METHOD 1631 AND 1638 INTERLABORATORY METHOD VALIDATION STUDY DATA

Table 7 compares detection and quantitation limits computed from data generated in the Method 1631 and Method 1638 interlaboratory studies. MDLs and MLs are those listed in EPA Methods 1631 and 1638. EPA computed IDEs and IQEs for the purpose of preparing this report. IDEs and IQEs computed by EPRI are from the EPRI reports on EPA Method 1631 and Method 1638 studies.

In reviewing these data it must be recognized that the EPA MDLs and MLs are the result of selecting the highest MDL in EPA's single-laboratory studies or among MDLs from the interlaboratory study, whereas the IDEs and IQEs are the result of a statistical process that includes recovery correction, correction for bias in the sample standard deviation (IQE only), allowance for prediction and tolerance intervals, interlaboratory variability, and model selection. The most significant reason for the instances of a large disparity between the EPA-determined IDEs/IQEs and the EPRI-determined IDEs/IQEs is model selection. EPA selected the model based on a strict application of the IDE and IQE procedures by a senior statistician. For those instances in which EPA and EPRI selected the same model, the IDEs and IQEs are nearly the same.

Table 8 compares IDEs resulting from the four main model types described in the ASTM IDE and IQE procedures. IDEs resulting from the constant model were the highest for all analytes. IDEs resulting from the other three models were almost equal for some analytes (lead, for example), and differed by more than an order of magnitude for others (mercury, for example). For two analytes, the IDE estimated using the linear model was negative. This was due to a negative intercept estimate in the precision model. The ASTM IDE and IQE procedures dictate that the linear model should not be used in this situation. No IDE could be calculated using the hybrid model for silver, because the IDE did not

converge to a single value using the calculated model for precision. This failure to converge is consistent with results for this analyte presented by EPRI.

DISCUSSION

Negative detection limits for the ISO/IUPAC MDV

The calculated MDV was negative for 26 analytes in the Episode 6000 data. Negative MDVs are attributable to the use of a regression model to estimate recovery at each concentration. The standard errors and correlation of the regression parameters are included in the calculation of the MDV. Analytes for which the MDV was negative seemed to coincide with an unusually large standard error of the regression intercept, which generally occurred when the estimated intercept was strongly negative. The large standard error of the intercept was likely due to extrapolating the recovery model to zero concentration; the error around a regression line is greatest for concentrations furthest away from the mean spike level. The effect of this extrapolation may also be seen in the Episode 6000 data. No negative results were used in the MDV and LOQ calculations, yet the median recovery intercept for these analytes was equal to -0.10. The standard errors of the intercept and slope estimates were generally high (intercept median=0.25, slope median=0.010), and therefore the estimated intercept and slope terms were frequently not significantly different from 0 and 1, respectively (intercept: not different from zero for 166 analytes/methods; slope not significantly different from 1 for 106 analytes; both intercept and slope not significant for 78 analytes). Because the recovery model parameters are not significant for the majority of analytes, and both the estimated slope and the standard errors of the slope and intercept are included in the calculation of the MDV and LOQ, the inclusion of the recovery model estimates may bias the calculated limits, to the point that the resulting MDV can be negative.

Effect of number and spacing of concentrations for determination of the SL-IDE and SL-IQE

Tests in the Episode 6000 studies were conducted at 16 concentration levels. The IDE procedure suggests 5 concentration levels. Based on statistical theory we would expect the number and spacing of concentration levels to affect the outcome, with a larger number of concentrations producing a more reliable estimate. We used the Episode 6000 data set to test this hypothesis.

The IDE procedure suggests spike concentrations at 0.5, 1.0, 2, 4, and 8 times an initial estimate of the IDE (IDE_0). IDE_0 is estimated at 10 times the standard deviation of replicates of a blank or the lowest level that can be measured. EPA's Episode 6000 database contain results of analysis of 7 replicates at 16 concentration levels from 0.1 to 100 times the initial estimate of the MDL (a factor of 1000). Between 0.1 and 10 times the MDL, the data are spaced a factor of approximately 1.5 apart. Above 10 times the MDL, the data are spaced at 10, 20, 50 and 100 times the MDL. The reason for the narrow spacing between 0.1 to 10 times the MDL was to attempt to allow more precise characterization of variability in the region of the MDL.

The SL-IDEs and SL-IQEs in Tables 2 and 4, respectively, were computed and reported using all 16 concentration levels because data were available at all of these levels. However, to determine the effect of the IDE procedure, a separate data analysis was performed. In this separate analysis, concentration levels were limited to a total of 5, and the 5 levels were selected to be as consistent as possible with the levels specified in the IDE procedure; i.e., at 5, 10, 20, 40, and 80 times the standard deviation of replicate measurements of a blank or the lowest level at which measurements could be made. The statement "lowest level at which measurements can be made" can be interpreted to mean inclusion or exclusion of results containing zeros and/or negative numbers. For purposes of the evaluation,

concentrations that produced results containing a zero or negative number were excluded; i.e., the lowest concentration that contained no zeros or negative numbers was chosen as the concentration at which the standard deviation would be calculated for the purpose of estimating IDE_0 and IQE_0 . Although some statisticians may argue that zeros and negative numbers should be included, nearly all analytical chemists would eliminate such numbers because they have no physical meaning.

The SL-IDE was calculated after selecting the levels based on IDE_0 and the results were compared to results produced when all 16 levels were included in calculating the SL-IDE. Results are summarized in Table 9. This table shows that the median 16-point IDE is approximately 1.3 times greater than the median 5-point IDE. For those instances in which the same model was chosen (108 out of 198), the median 16-point IDE was approximately 1.4 times higher than the median 5-point IDE, which was significantly different from 1.0 based on a sign test ($p < 0.0001$). For those instances in which a different model was chosen (90 out of 198), the median 16-point IDE was approximately 0.9 times the median 5-point IDE, which was not significantly different from 1.0 ($p = 0.83$). Because the choice of model can have a confounding effect on any differences between 16-point and 5-point SL-IDEs, the focus should be on the instances in which the same model was chosen. For these instances, the results indicate that only data in the region of detection and quantitation should be used to establish a detection or quantitation limit.

Parallel reasoning can be applied to the IQE, because IQE_0 is specified and the IQE is developed in a way analogous to that for the IDE.

RSD AT THE ML IN THE EPISODE 6000 STUDY

The minimum level of quantitation (ML) is directed at the level at which 10% relative standard deviation (RSD) is attained. However, because the ML is not established at exactly 10% RSD, but is determined by multiplying the standard deviation that is obtained in determination of an MDL by 10 (as recommended by both ACS and Currie for ACS and ISO/IUPAC LOQs), the resulting RSD may not be 10%. The Episode 6000 data provided the opportunity to determine the actual value of the RSD at the ML. Results of the determination showed that the overall median RSD at the ML across all 198 analytes in the Episode 6000 study was 7 %, and the median RSD per analytical technique ranged between 6 and 16 percent by analytical technique for analytes in the 10 analytical techniques in the study. For 126 of the 198 individual analytes, the RSD fell between 5% and 15%. For the majority of the analytes that fell outside this range (56 out of 72), the RSD was less than 5%.

There was a spike concentration at the ML for approximately 80 percent of the analytes in the Episode 6000 study so the RSD could be determined directly for these analytes. For a few analytes there was not a spike concentration at the ML, and the RSD was determined by interpolation between spike levels for these analytes. However, for 82 out of 198 analytes, the concentration at the ML was below the range of the data; i.e., below the lowest spike concentration that returned non-zero and non-negative results. These instances were for organic analytes determined by EPA Methods 502.2 and 524.2. The reason that these low values occur is because hardware and/or software thresholds in chromatographic instruments that are set to eliminate spurious noise signals also filter out responses at low concentrations. For instances in which this occurred, the ML was calculated as the lowest concentration at which non-zero and non-negative results were not reported, and the RSD was calculated at this concentration. The median RSD for these analytes was 5%, compared to a median of 9% where this did not happen.

CONCLUSIONS

The comparisons of detection and quantitation limits show high variability among the limits calculated using the different concepts, even with data containing 7 replicates at 16 concentration levels (see the summary statistics at the end of Tables 3 and 5, Table 6, and the final conclusion below). The net effect is that the systematic differences among detection and quantitation limits produced by the various concepts are overwhelmed by variability; i.e., there is a small systematic difference among the concepts but great variability in the detection and quantitation limits for a given analyte. This result is not surprising given the variability of data in the region of detection and quantitation. However, it is difficult to postulate a solution to the problem. Gathering more data in the region of detection and quantitation would appear to be a solution, but 91 data points were gathered for each analyte in the region between 0.1 and 10 times the MDL in the Episode 6000 studies, and it is unlikely that any organization could afford to gather even this amount of data for determination of a detection limit. Given the high degree of variability of the data, EPA's approach of conducting single-laboratory study to gain a first estimate, then multiple single-laboratory studies to verify or revise the estimate, then an interlaboratory study, where warranted, to further verify and revise the estimate, is a reasonable means of establishing detection and quantitation limits because of the checks and balances that occur at each step.

A second conclusion is that using a regression line to estimate a recovery correction at zero concentration causes great swings in the resulting detection and quantitation limits such as the ISO/IUPAC MDV and LOQ. The estimated regression parameters for the recovery models were often not significant, and the inclusion of the estimated slope and the standard errors of the slope and intercept will therefore unnecessarily bias the calculated MDV and LOQ, such that the calculated MDVs may be negative (see Discussion section "Negative detection limits for the ISO/IUPAC MDV, and Table 2 for instances of negative detection limits"). The estimated recovery model used in calculating the IDE and IQE is also strongly affected by the chosen model of variability vs. concentration (see Table 8). Even though a linear regression is used to model recovery in each case, the weights used in the model are calculated based on the variability model, and can vary greatly when the number of concentrations used is low. For the Episode 6000 data, the median RSD of the recovery slopes from the four different models used in the IDE calculations for a given analyte and method was 5%. In addition, for 75 of the analytes and methods (38%), at least one estimated recovery slope was greater than 1, and at least one was less than 1. This suggests that the method could be considered to be high biased (and the final IDE and IQE would be decreased by the recovery correction) and low biased (and the final IDE and IQE increased) for the analyte, depending on the chosen precision model. For many analytes the slopes were not significantly different from 1, suggesting that a recovery correction may not be appropriate at all. This is in addition to the philosophical issue as to whether recovery correction is warranted. If there is to be a correction for recovery, it may be better to use some average or median value than a regression, or use a measured value near the region of interest.

A third conclusion is that further work will need to be done on the ASTM IDE and IQE before they can be used routinely, not only because of the complexity of the procedures, but also because of the ambiguity in determining that the correct model has been selected. (For the consequences of model selection, compare the IDEs and IQEs determined by EPA and EPRI in Table 7, and the IDEs calculated from the different model types in Table 8. Some differ considerably as a result of model selection in application of the IDE and IQE procedures by different statisticians.)

A fourth conclusion is that quantitation limit concepts such as EPA's ML and the ACS and ISO/IUPAC LOQ that are directed 10% RSD actually produce RSDs that are in the range of the 10% intended (see the discussion in the Section titled "RSD at the ML in the Episode 6000 Study"). The median RSDs for each method in the Episode 6000 data set ranged from 6% to 16%, and 64% of the individual analyte RSDs fell between 5% and 15%.

Finally, a statement needs to be made about the overall philosophy behind the concepts. A natural progression through time is to refine any theory or concept. However, in the case of detection and quantitation in analytical chemistry (and as applied to other disciplines), the effort to further refine concepts may be futile, given the variability of the data that must be used with any concept. However, it is clear no concept produces the “right” answer. Different concepts allow for different sources of variability resulting in shifts in the magnitude of the limit. For the EPA, ACS, ISO/IUPAC, and ASTM concepts, the most significant causes of this shift appear to be 1) the false positive rate ($\alpha = 0.01$ or 0.05), 2) the allowance for false negatives ($\beta = 0.05$), 3) recovery correction, 4) correction for the bias in the calculated standard deviation (IQE only), 5) interlaboratory variability, and 6) model selection in the IDE and IQE.

Table 1. Data Sets Suggested by Petitioners

| Data Set and Year | Analyte and Technology |
|--------------------------|-------------------------------|
| AAMA 1996-7 | Metals by ICP/AES (200.7) |
| AAMA 1996-7 | Mercury by CVAA (245.2) |
| AAMA 1996-7 | PCBs by GC/ECD (608.2) |
| MMA 2000-1 | PCB 1216 and 1260 by GC/ECD |
| EPA/EPRI 1997-8 | Mercury by CVAF (1631) |
| EPA/EPRI 1997-8 | Metals by ICP/MS (1638) |
| EPRI 1987 | Metals by GFAA (EPA 200) |
| EPRI 1990 | Metals by ICP/AES (EPA 200.7) |
| EPRI 1994 | Al, Be, Ti by GFAA (EPA 200) |
| EPRI 1996 | Cd, As, Cr by GFAA (EPA 200) |

**Table 2. Comparison of Detection Limits ($\mu\text{g/L}$ except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|-----------------------------|--------|-----------|----------------|------------|------------|----------------|
| 1,1,1,2-tetrachloroethane | 502.2 | ELCD | 0.041 | 0.028 | 0.054 | 0.028 |
| 1,1,1,2-tetrachloroethane | 524.2 | | 0.052 | 0.039 | -0.030 | 0.206 |
| 1,1,1-trichloroethane | 502.2 | ELCD | 0.012 | 0.009 | 0.017 | 0.035 |
| 1,1,1-trichloroethane | 524.2 | | 0.055 | 0.021 | 0.007 | 0.268 |
| 1,1,2,2-tce+1,2,3-tcp | 502.2 | ELCD | 0.064 | 0.225 | 0.417 | 0.170 |
| 1,1,2,2-tetrachloroethane | 524.2 | | 0.132 | 0.131 | 0.139 | 0.377 |
| 1,1,2-trichloroethane | 502.2 | ELCD | 0.024 | 0.055 | 0.103 | 0.026 |
| 1,1,2-trichloroethane | 524.2 | | 0.075 | 0.043 | 0.045 | 0.284 |
| 1,1-dichloroethane | 502.2 | ELCD | 0.010 | 0.008 | 0.016 | 0.066 |
| 1,1-dichloroethane | 524.2 | | 0.033 | 0.020 | 0.018 | 0.206 |
| 1,1-dichloroethene | 502.2 | ELCD | 0.038 | 0.013 | 0.032 | 0.193 |
| 1,1-dichloroethene | 524.2 | | 0.054 | 0.035 | -0.030 | 0.278 |
| 1,1-dichloropropanone | 524.2 | | 5.184 | 3.146 | 5.657 | 6.032 |
| 1,1-dichloropropene | 524.2 | | 0.045 | 0.012 | -0.020 | 0.247 |
| 1,2,3-trichlorobenzene | 502.2 | ELCD | 0.048 | 0.308 | 0.599 | 0.122 |
| 1,2,3-trichlorobenzene | 502.2 | PID | 0.057 | 0.301 | 0.623 | 0.114 |
| 1,2,3-trichlorobenzene | 524.2 | | 0.070 | 0.040 | 0.036 | 0.259 |
| 1,2,3-trichloropropane | 524.2 | | 7.328 | 0.046 | 0.042 | 1.206 |
| 1,2,4-trichlorobenzene | 502.2 | ELCD | 0.022 | 0.189 | 0.393 | 0.077 |
| 1,2,4-trichlorobenzene | 502.2 | PID | 0.070 | 0.221 | 0.471 | 0.124 |
| 1,2,4-trichlorobenzene | 524.2 | | 0.053 | 0.050 | 0.057 | 0.208 |
| 1,2,4-trimethylbenzene | 502.2 | PID | 0.095 | 0.075 | 0.167 | 0.123 |
| 1,2,4-trimethylbenzene | 524.2 | | 0.012 | 0.009 | 0.017 | 0.129 |
| 1,2-dibromo-3-chloropropane | 524.2 | | 1.457 | 0.391 | 0.702 | 1.619 |
| 1,2-dibromoethane | 502.2 | ELCD | 0.096 | 0.028 | 0.056 | 0.143 |
| 1,2-dibromoethane | 524.2 | | 0.127 | 0.117 | 0.175 | 0.289 |
| 1,2-dichlorobenzene | 502.2 | ELCD | 0.035 | 0.073 | 0.144 | 0.053 |
| 1,2-dichlorobenzene | 502.2 | PID | 0.033 | 0.024 | 0.054 | 0.147 |
| 1,2-dichlorobenzene | 524.2 | | 0.030 | 0.023 | -0.010 | 0.112 |
| 1,2-dichloroethane | 502.2 | ELCD | 0.017 | 0.017 | 0.032 | 0.037 |
| 1,2-dichloroethane | 524.2 | | 0.039 | 0.024 | 0.017 | 0.229 |
| 1,2-dichloropropane | 502.2 | ELCD | 0.023 | 0.196 | 0.393 | 0.037 |
| 1,2-dichloropropane | 524.2 | | 0.056 | 0.031 | 0.029 | 0.221 |
| 1,3,5-tmb+4-chlorotoluene | 502.2 | PID | 0.067 | 0.201 | 0.449 | 0.108 |

**Table 2. Comparison of Detection Limits (µg/L except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|----------------------------------|--------|-----------|----------------|------------|------------|----------------|
| 1,3,5-trimethylbenzene | 524.2 | | 0.011 | 0.008 | 0.009 | 0.117 |
| 1,3-dichlorobenzene | 502.2 | ELCD | 0.035 | 0.048 | 0.104 | 0.100 |
| 1,3-dichlorobenzene | 502.2 | PID | 0.093 | 0.134 | 0.297 | 0.123 |
| 1,3-dichlorobenzene | 524.2 | | 0.023 | 0.016 | -0.010 | 0.126 |
| 1,3-dichloropropane | 502.2 | ELCD | 0.016 | 0.071 | 0.135 | 0.045 |
| 1,3-dichloropropane | 524.2 | | 0.038 | 0.024 | -0.010 | 0.170 |
| 1,4-dichlorobenzene | 502.2 | ELCD | 0.026 | 0.101 | 0.198 | 0.054 |
| 1,4-dichlorobenzene | 524.2 | | 0.023 | 0.017 | -0.040 | 0.110 |
| 1-chlorobutane | 524.2 | | 0.020 | 0.016 | 0.019 | 0.200 |
| 2,2-dichloropropane | 524.2 | | 2.376 | 0.103 | -0.120 | 0.700 |
| 2-butanone | 524.2 | | 0.417 | 0.297 | 0.515 | 0.775 |
| 2-chlorotoluene | 502.2 | ELCD | 0.108 | 0.162 | 0.318 | 0.184 |
| 2-chlorotoluene | 502.2 | PID | 0.238 | 0.454 | 1.022 | 0.222 |
| 2-chlorotoluene | 524.2 | | 0.016 | 0.009 | 0.003 | 0.121 |
| 2-hexanone | 524.2 | | 1.316 | 0.148 | 0.234 | 0.815 |
| 2-nitropropane | 524.2 | | 0.901 | 0.275 | 0.457 | 0.965 |
| 4-chlorotoluene | 502.2 | ELCD | 0.110 | 0.127 | 0.234 | 0.159 |
| 4-chlorotoluene | 524.2 | | 0.010 | 0.008 | 0.008 | 0.102 |
| 4-isopropyltoluene | 524.2 | | 0.010 | 0.008 | 0.005 | 0.102 |
| 4-methyl-2-pentanone | 524.2 | | 0.812 | 0.480 | 0.740 | 1.060 |
| Acetone | 524.2 | | 0.859 | 0.440 | 0.806 | 2.025 |
| Acrylonitrile | 524.2 | | 0.863 | 0.444 | 0.658 | 1.197 |
| Allyl Chloride | 524.2 | | 0.032 | 0.026 | 0.010 | 0.203 |
| Aluminum | 1620 | | 29.555 | 15.043 | 28.670 | 198.565 |
| Aluminum | 200.8 | | 19.145 | 1.690 | 3.547 | 12.004 |
| Ammonia as Nitrogen ¹ | 350.3 | | 0.010 | 0.007 | 0.014 | 0.013 |
| Antimony | 1620 | | 1.552 | 0.801 | 1.754 | 4.087 |
| Antimony | 200.8 | | 0.178 | 0.003 | 0.007 | 0.018 |
| Arsenic | 1620 | | 1.065 | 0.917 | 1.375 | 1.463 |
| Arsenic | 200.8 | | 0.226 | 0.137 | 0.272 | 0.346 |
| Barium | 1620 | | 1.702 | 1.337 | 1.831 | 1.762 |
| Barium | 200.8 | | 0.033 | 0.029 | 0.061 | 0.079 |
| Benzene | 502.2 | PID | 0.030 | 0.043 | 0.099 | 0.077 |
| Benzene | 524.2 | | 0.014 | 0.014 | 0.027 | 0.115 |

**Table 2. Comparison of Detection Limits (µg/L except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|------------------------|---------------|------------------|------------------------|--------------------|--------------------|------------------------|
| Beryllium | 1620 | | 0.528 | 0.339 | 0.408 | 0.428 |
| Beryllium | 200.8 | | 0.007 | 0.004 | 0.006 | 0.019 |
| Boron | 1620 | | 15.387 | 10.356 | 17.790 | 19.884 |
| Bromobenzene | 502.2 | ELCD | 0.131 | 0.093 | 0.187 | 0.729 |
| Bromobenzene | 502.2 | PID | 0.012 | 0.286 | 0.619 | 0.048 |
| Bromobenzene | 524.2 | | 0.044 | 0.036 | -0.050 | 0.175 |
| Bromochloromethane | 502.2 | ELCD | 0.013 | 0.016 | 0.032 | 0.462 |
| Bromochloromethane | 524.2 | | 0.125 | 0.113 | 0.165 | 0.309 |
| Bromodichloromethane | 502.2 | ELCD | 0.004 | 0.016 | 0.032 | 0.064 |
| Bromodichloromethane | 524.2 | | 0.043 | 0.026 | 0.023 | 0.182 |
| Bromoform | 502.2 | ELCD | 0.006 | 0.009 | 0.016 | 1.450 |
| Bromoform | 524.2 | | 0.123 | 0.065 | 0.040 | 0.350 |
| Bromomethane | 502.2 | ELCD | 0.267 | 0.047 | 0.010 | 6.993 |
| Bromomethane | 524.2 | | 0.068 | 0.055 | 0.060 | 0.238 |
| Cadmium | 1620 | | 0.127 | 0.079 | 0.134 | 0.184 |
| Cadmium | 200.8 | | 0.004 | 0.007 | 0.012 | 0.011 |
| Calcium | 1620 | | 36.726 | 35.822 | 72.400 | 39.651 |
| Carbon Disulfide | 524.2 | | 0.027 | 0.016 | -0.030 | 0.203 |
| Carbon Tetrachloride | 524.2 | | 0.038 | 0.027 | -0.030 | 0.258 |
| Carbontet+1,1-dcp | 502.2 | ELCD | 0.029 | 0.028 | 0.060 | 0.062 |
| Chloroacetonitrile | 524.2 | | 0.919 | 0.773 | 1.535 | 1.535 |
| Chlorobenzene | 502.2 | ELCD | 0.011 | 0.016 | 0.034 | 0.440 |
| Chlorobenzene | 502.2 | PID | 0.030 | 0.080 | 0.178 | 0.062 |
| Chlorobenzene | 524.2 | | 0.025 | 0.022 | 0.016 | 0.119 |
| Chloroethane | 502.2 | ELCD | 0.108 | 0.006 | 0.004 | 2.492 |
| Chloroethane | 524.2 | | 0.066 | 0.041 | 0.042 | 0.336 |
| Chloroform | 502.2 | ELCD | 0.043 | 0.410 | 0.758 | 0.031 |
| Chloroform | 524.2 | | 0.036 | 0.027 | 0.025 | 0.203 |
| Chloromethane | 502.2 | ELCD | 0.070 | 0.090 | 0.240 | 0.204 |
| Chloromethane | 524.2 | | 0.045 | 0.036 | 0.066 | 0.215 |
| Chromium | 1620 | | 0.310 | 0.254 | 0.386 | 0.478 |
| Chromium | 200.8 | | 0.073 | 0.062 | 0.125 | 0.393 |
| Cis-1,2-dce+2,2-dcp | 502.2 | ELCD | 0.013 | 0.017 | 0.034 | 0.043 |
| Cis-1,2-dichloroethene | 524.2 | | 0.040 | 0.033 | -0.010 | 0.203 |

**Table 2. Comparison of Detection Limits ($\mu\text{g/L}$ except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|------------------------------|--------|-----------|----------------|------------|------------|----------------|
| Cis-1,3-dichloropropene | 502.2 | ELCD | 0.007 | 0.016 | 0.031 | 0.059 |
| Cis-1,3-dichloropropene | 502.2 | PID | 0.057 | 0.048 | 0.099 | 0.077 |
| Cis-1,3-dichloropropene | 524.2 | | 0.038 | 0.024 | 0.001 | 0.158 |
| Cobalt | 1620 | | 9.820 | 4.017 | 8.094 | 15.560 |
| Cobalt | 200.8 | | 0.001 | 0.001 | -0.070 | 0.008 |
| Copper | 1620 | | 6.046 | 4.990 | 10.510 | 20.328 |
| Copper | 200.8 | | 0.037 | 0.027 | 0.053 | 0.770 |
| Dibromochloromethane | 502.2 | ELCD | 0.009 | 0.019 | 0.039 | 0.418 |
| Dibromochloromethane | 524.2 | | 0.051 | 0.031 | 0.011 | 0.253 |
| Dibromomethane | 502.2 | ELCD | 0.007 | 0.047 | 0.096 | 0.441 |
| Dibromomethane | 524.2 | | 0.102 | 0.082 | 0.117 | 0.342 |
| Dichlorodifluoromethane | 502.2 | ELCD | 0.009 | 1.453 | 1.715 | 0.087 |
| Dichlorodifluoromethane | 524.2 | | 0.083 | 0.054 | 0.046 | 0.420 |
| Diethyl Ether | 524.2 | | 0.120 | 0.114 | 0.169 | 0.340 |
| Ethyl Methacrylate | 524.2 | | 0.045 | 0.031 | 0.018 | 0.244 |
| Ethylbenzene | 502.2 | PID | 0.021 | 0.053 | 0.122 | 0.075 |
| Ethylbenzene | 524.2 | | 0.033 | 0.028 | -0.020 | 0.167 |
| Hardness ¹ | 130.2 | | 0.828 | 0.555 | 1.152 | 2.152 |
| Hexachlorobutadiene | 502.2 | ELCD | 0.043 | 0.240 | 0.502 | 0.090 |
| Hexachlorobutadiene | 524.2 | | 0.068 | 0.035 | -0.020 | 0.263 |
| Hexachloroethane | 524.2 | | 0.056 | 0.049 | 0.044 | 0.234 |
| Hexchlobutadiene+naphthalene | 502.2 | PID | 0.649 | 0.924 | 2.083 | 0.598 |
| Iron | 1620 | | 90.409 | 270.433 | 472.200 | 345.686 |
| Isopropylbenzene | 502.2 | PID | 0.020 | 0.051 | 0.120 | 0.059 |
| Isopropylbenzene | 524.2 | | 0.011 | 0.010 | 0.012 | 0.113 |
| Lead | 1620 | | 1.647 | 1.186 | 1.965 | 2.317 |
| Lead | 200.8 | | 0.655 | 0.061 | 0.120 | 0.197 |
| M+p Xylene | 502.2 | PID | 0.090 | 0.099 | 0.225 | 0.116 |
| M+p Xylene | 524.2 | | 0.013 | 0.008 | 0.006 | 0.127 |
| Magnesium | 1620 | | 103.033 | 88.729 | 175.300 | 99.662 |
| Manganese | 1620 | | 6.856 | 1.081 | 2.591 | 6.531 |
| Manganese | 200.8 | | 0.031 | 0.030 | 0.049 | 0.106 |
| Mercury | 200.8 | | 0.004 | 0.003 | -0.020 | 0.062 |
| Methacrylonitrile | 524.2 | | 0.356 | 0.228 | 0.368 | 0.643 |

**Table 2. Comparison of Detection Limits ($\mu\text{g/L}$ except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|-------------------------|--------|-----------|----------------|------------|------------|----------------|
| Methyl Iodide | 524.2 | | 0.025 | 0.023 | -0.010 | 0.173 |
| Methyl Tert-butyl Ether | 524.2 | | 0.026 | 0.016 | -0.030 | 0.195 |
| Methylacrylate | 524.2 | | 0.220 | 0.202 | 0.356 | 0.549 |
| Methylene Chloride | 502.2 | ELCD | 0.128 | 1.835 | 5.018 | 2.727 |
| Methylene Chloride | 524.2 | | 0.082 | 0.072 | 0.098 | 0.276 |
| Methylmethacrylate | 524.2 | | 0.225 | 0.085 | 0.120 | 0.484 |
| Molybdenum | 1620 | | 2.455 | 1.714 | 3.787 | 2.917 |
| Molybdenum | 200.8 | | 0.004 | 0.003 | 0.000 | 0.262 |
| N-butylbenzene | 502.2 | PID | 0.030 | 0.069 | 0.151 | 0.139 |
| N-butylbenzene | 524.2 | | 0.016 | 0.014 | 0.027 | 0.136 |
| N-propylbenzene | 502.2 | PID | 0.040 | 0.597 | 1.340 | 0.089 |
| N-propylbenzene | 524.2 | | 0.038 | 0.026 | -0.040 | 0.231 |
| Naphthalene | 524.2 | | 0.048 | 0.040 | 0.047 | 0.175 |
| Nickel | 1620 | | 20.219 | 13.262 | 25.700 | 23.784 |
| Nickel | 200.8 | | 0.146 | 0.058 | 0.107 | 0.076 |
| o-xylene | 524.2 | | 0.018 | 0.015 | -0.030 | 0.161 |
| o-xylene+styrene | 502.2 | PID | 0.059 | 0.118 | 0.263 | 0.111 |
| P-isoproptol+1,4-dcb | 502.2 | PID | 0.073 | 0.152 | 0.332 | 0.153 |
| Pentachloroethane | 524.2 | | 0.553 | 0.019 | -0.080 | 0.337 |
| Sec-butylbenzene | 502.2 | PID | 0.055 | 0.058 | 0.133 | 0.079 |
| Sec-butylbenzene | 524.2 | | 0.014 | 0.011 | -0.010 | 0.132 |
| Selenium | 1620 | | 0.849 | 0.619 | 1.493 | 1.915 |
| Selenium | 200.8 | | 0.192 | 0.156 | 0.302 | 0.410 |
| Silver | 1620 | | 4.907 | 3.588 | 6.495 | 10.219 |
| Silver | 200.8 | | 0.004 | 0.002 | 0.004 | 0.010 |
| Sodium | 1620 | | 69.530 | 49.595 | 97.650 | 133.007 |
| Styrene | 524.2 | | 0.014 | 0.011 | 0.011 | 0.119 |
| Tert-butylbenzene | 502.2 | PID | 0.029 | 0.058 | 0.137 | 0.073 |
| Tert-butylbenzene | 524.2 | | 0.022 | 0.012 | 0.023 | 0.170 |
| Tetrachloroethene | 502.2 | ELCD | 0.018 | 0.200 | 0.429 | 0.051 |
| Tetrachloroethene | 502.2 | PID | 0.062 | 0.319 | 0.753 | 0.157 |
| Tetrachloroethene | 524.2 | | 0.085 | 0.084 | 0.058 | 0.379 |
| Thallium | 1620 | | 0.512 | 0.651 | 1.406 | 1.208 |
| Thallium | 200.8 | | 0.000 | 0.000 | 0.001 | 0.001 |

**Table 2. Comparison of Detection Limits ($\mu\text{g/L}$ except where footnoted)
for the Episode 6000 Data set**

| Analyte | Method | Procedure | EPA/ ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
|-------------------------------------|--------|-----------|----------------|------------|------------|----------------|
| Thorium | 200.8 | | 0.001 | 0.001 | 0.000 | 0.003 |
| Tin | 1620 | | 3.670 | 2.019 | 3.143 | 1.200 |
| Titanium | 1620 | | 4.777 | 4.453 | 8.050 | 5.238 |
| Toluene | 502.2 | PID | 0.070 | 0.064 | 0.145 | 0.061 |
| Toluene | 524.2 | | 0.020 | 0.006 | 0.000 | 0.130 |
| Total Phosphorus ¹ | 365.2 | | 0.006 | 0.005 | 0.009 | 0.013 |
| Total Suspended Solids ¹ | 160.2 | | 1.170 | 0.948 | 1.945 | 2.877 |
| Trans-1,2-dichloroethene | 502.2 | ELCD | 0.041 | 0.174 | 0.382 | 0.065 |
| Trans-1,2-dichloroethene | 524.2 | | 0.038 | 0.032 | -0.010 | 0.255 |
| Trans-1,3-dichloropropene | 502.2 | ELCD | 0.012 | 0.013 | 0.026 | 0.082 |
| Trans-1,3-dichloropropene | 502.2 | PID | 0.058 | 0.037 | 0.079 | 0.085 |
| Trans-1,3-dichloropropene | 524.2 | | 0.051 | 0.025 | 0.000 | 0.204 |
| Trans-1,4-dichloro-2-butene | 524.2 | | 0.512 | 0.348 | 0.589 | 1.182 |
| Trichloroethene | 502.2 | ELCD | 0.012 | 0.014 | 0.029 | 0.050 |
| Trichloroethene | 502.2 | PID | 0.027 | 0.043 | 0.098 | 0.096 |
| Trichloroethene | 524.2 | | 0.061 | 0.058 | 0.062 | 0.288 |
| Trichlorofluoromethane | 502.2 | ELCD | 0.108 | 0.012 | 0.028 | 1.997 |
| Trichlorofluoromethane | 524.2 | | 0.087 | 0.075 | 0.046 | 0.307 |
| Uranium | 200.8 | | 0.000 | 0.000 | 0.000 | 0.001 |
| Vanadium | 1620 | | 7.344 | 4.207 | 8.359 | 10.063 |
| Vanadium | 200.8 | | 0.555 | 0.512 | 0.994 | 0.845 |
| Vinyl Chloride | 502.2 | ELCD | 0.270 | 0.039 | 0.072 | 3.521 |
| Vinyl Chloride | 524.2 | | 0.043 | 0.031 | 0.000 | 0.295 |
| WAD Cyanide | 1677 | | 0.572 | 0.169 | 0.319 | 0.672 |
| Xylene (Total) | 524.2 | | 0.009 | 0.005 | 0.008 | 0.111 |
| Yttrium | 1620 | | 1.923 | 1.370 | 2.518 | 3.119 |
| Zinc | 1620 | | 2.597 | 2.301 | 3.697 | 4.415 |
| Zinc | 200.8 | | 0.900 | 0.461 | 0.806 | 1.497 |

¹ Results reported as mg/L

Note: ELCD or PID in the Procedure column indicates the photo-ionization detector (PID) or electrolytic conductivity detector (ELCD) in EPA Method 502.2

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|-----------------------------|--------|-----------|--------------------|--------------------|----------------|
| 1,1,1,2-tetrachloroethane | 502.2 | ELCD | 0.691 | 1.313 | 0.679 |
| 1,1,1,2-tetrachloroethane | 524.2 | | 0.747 | -0.645 | 3.953 |
| 1,1,1-trichloroethane | 502.2 | ELCD | 0.707 | 1.387 | 2.861 |
| 1,1,1-trichloroethane | 524.2 | | 0.382 | 0.135 | 4.843 |
| 1,1,2,2-tce+1,2,3-tcp | 502.2 | ELCD | 3.522 | 6.531 | 2.664 |
| 1,1,2,2-tetrachloroethane | 524.2 | | 0.994 | 1.054 | 2.853 |
| 1,1,2-trichloroethane | 502.2 | ELCD | 2.266 | 4.207 | 1.062 |
| 1,1,2-trichloroethane | 524.2 | | 0.580 | 0.602 | 3.796 |
| 1,1-dichloroethane | 502.2 | ELCD | 0.786 | 1.645 | 6.600 |
| 1,1-dichloroethane | 524.2 | | 0.596 | 0.561 | 6.281 |
| 1,1-dichloroethene | 502.2 | ELCD | 0.348 | 0.845 | 5.018 |
| 1,1-dichloroethene | 524.2 | | 0.647 | -0.477 | 5.102 |
| 1,1-dichloropropanone | 524.2 | | 0.607 | 1.091 | 1.164 |
| 1,1-dichloropropene | 524.2 | | 0.261 | -0.536 | 5.525 |
| 1,2,3-trichlorobenzene | 502.2 | ELCD | 6.454 | 12.543 | 2.548 |
| 1,2,3-trichlorobenzene | 502.2 | PID | 5.268 | 10.902 | 1.988 |
| 1,2,3-trichlorobenzene | 524.2 | | 0.578 | 0.517 | 3.706 |
| 1,2,3-trichloropropane | 524.2 | | 0.006 | 0.006 | 0.165 |
| 1,2,4-trichlorobenzene | 502.2 | ELCD | 8.731 | 18.213 | 3.548 |
| 1,2,4-trichlorobenzene | 502.2 | PID | 3.143 | 6.704 | 1.773 |
| 1,2,4-trichlorobenzene | 524.2 | | 0.951 | 1.069 | 3.921 |
| 1,2,4-trimethylbenzene | 502.2 | PID | 0.792 | 1.766 | 1.297 |
| 1,2,4-trimethylbenzene | 524.2 | | 0.772 | 1.419 | 10.509 |
| 1,2-dibromo-3-chloropropane | 524.2 | | 0.268 | 0.482 | 1.111 |
| 1,2-dibromoethane | 502.2 | ELCD | 0.297 | 0.582 | 1.496 |
| 1,2-dibromoethane | 524.2 | | 0.918 | 1.378 | 2.275 |
| 1,2-dichlorobenzene | 502.2 | ELCD | 2.079 | 4.106 | 1.525 |
| 1,2-dichlorobenzene | 502.2 | PID | 0.734 | 1.649 | 4.514 |
| 1,2-dichlorobenzene | 524.2 | | 0.755 | -0.319 | 3.729 |
| 1,2-dichloroethane | 502.2 | ELCD | 0.981 | 1.834 | 2.107 |
| 1,2-dichloroethane | 524.2 | | 0.609 | 0.445 | 5.882 |
| 1,2-dichloropropane | 502.2 | ELCD | 8.543 | 17.176 | 1.602 |
| 1,2-dichloropropane | 524.2 | | 0.540 | 0.519 | 3.915 |
| 1,3,5-tmb+4-chlorotoluene | 502.2 | PID | 2.984 | 6.658 | 1.598 |

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|------------------------|--------|-----------|-----------------|-----------------|----------------|
| 1,3,5-trimethylbenzene | 524.2 | | 0.710 | 0.843 | 10.494 |
| 1,3-dichlorobenzene | 502.2 | ELCD | 1.375 | 2.980 | 2.850 |
| 1,3-dichlorobenzene | 502.2 | PID | 1.445 | 3.192 | 1.325 |
| 1,3-dichlorobenzene | 524.2 | | 0.699 | -0.357 | 5.418 |
| 1,3-dichloropropane | 502.2 | ELCD | 4.518 | 8.588 | 2.839 |
| 1,3-dichloropropane | 524.2 | | 0.628 | -0.212 | 4.457 |
| 1,4-dichlorobenzene | 502.2 | ELCD | 3.876 | 7.630 | 2.068 |
| 1,4-dichlorobenzene | 524.2 | | 0.714 | -1.534 | 4.699 |
| 1-chlorobutane | 524.2 | | 0.786 | 0.974 | 10.051 |
| 2,2-dichloropropane | 524.2 | | 0.044 | -0.049 | 0.294 |
| 2-butanone | 524.2 | | 0.713 | 1.235 | 1.860 |
| 2-chlorotoluene | 502.2 | ELCD | 1.498 | 2.947 | 1.705 |
| 2-chlorotoluene | 502.2 | PID | 1.907 | 4.290 | 0.931 |
| 2-chlorotoluene | 524.2 | | 0.570 | 0.215 | 7.599 |
| 2-hexanone | 524.2 | | 0.112 | 0.178 | 0.620 |
| 2-nitropropane | 524.2 | | 0.305 | 0.507 | 1.071 |
| 4-chlorotoluene | 502.2 | ELCD | 1.152 | 2.135 | 1.445 |
| 4-chlorotoluene | 524.2 | | 0.803 | 0.849 | 10.563 |
| 4-isopropyltoluene | 524.2 | | 0.833 | 0.519 | 10.452 |
| 4-methyl-2-pentanone | 524.2 | | 0.591 | 0.911 | 1.305 |
| Acetone | 524.2 | | 0.512 | 0.938 | 2.358 |
| Acrylonitrile | 524.2 | | 0.515 | 0.763 | 1.387 |
| Allyl Chloride | 524.2 | | 0.820 | 0.306 | 6.297 |
| Aluminum | 1620 | | 0.509 | 0.970 | 6.718 |
| Aluminum | 200.8 | | 0.088 | 0.185 | 0.627 |
| Ammonia as Nitrogen | 350.3 | | 0.668 | 1.358 | 1.229 |
| Antimony | 1620 | | 0.516 | 1.130 | 2.634 |
| Antimony | 200.8 | | 0.018 | 0.037 | 0.102 |
| Arsenic | 1620 | | 0.861 | 1.290 | 1.373 |
| Arsenic | 200.8 | | 0.606 | 1.206 | 1.532 |
| Barium | 1620 | | 0.786 | 1.076 | 1.035 |
| Barium | 200.8 | | 0.885 | 1.855 | 2.421 |
| Benzene | 502.2 | PID | 1.435 | 3.283 | 2.537 |
| Benzene | 524.2 | | 0.981 | 1.839 | 7.986 |

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|------------------------|--------|-----------|-----------------|-----------------|----------------|
| Beryllium | 1620 | | 0.641 | 0.773 | 0.810 |
| Beryllium | 200.8 | | 0.564 | 0.846 | 2.667 |
| Boron | 1620 | | 0.673 | 1.156 | 1.292 |
| Bromobenzene | 502.2 | ELCD | 0.711 | 1.428 | 5.575 |
| Bromobenzene | 502.2 | PID | 23.364 | 50.631 | 3.947 |
| Bromobenzene | 524.2 | | 0.834 | -1.041 | 4.021 |
| Bromochloromethane | 502.2 | ELCD | 1.166 | 2.381 | 34.635 |
| Bromochloromethane | 524.2 | | 0.902 | 1.319 | 2.468 |
| Bromodichloromethane | 502.2 | ELCD | 3.804 | 7.753 | 15.209 |
| Bromodichloromethane | 524.2 | | 0.614 | 0.538 | 4.229 |
| Bromoform | 502.2 | ELCD | 1.503 | 2.627 | 242.497 |
| Bromoform | 524.2 | | 0.523 | 0.323 | 2.835 |
| Bromomethane | 502.2 | ELCD | 0.178 | 0.037 | 26.202 |
| Bromomethane | 524.2 | | 0.810 | 0.893 | 3.516 |
| Cadmium | 1620 | | 0.619 | 1.056 | 1.449 |
| Cadmium | 200.8 | | 1.769 | 2.985 | 2.793 |
| Calcium | 1620 | | 0.975 | 1.971 | 1.080 |
| Carbon Disulfide | 524.2 | | 0.582 | -1.199 | 7.610 |
| Carbon Tetrachloride | 524.2 | | 0.711 | -0.770 | 6.751 |
| Carbontet+1,1-dcp | 502.2 | ELCD | 0.986 | 2.089 | 2.162 |
| Chloroacetonitrile | 524.2 | | 0.841 | 1.671 | 1.671 |
| Chlorobenzene | 502.2 | ELCD | 1.373 | 2.964 | 38.584 |
| Chlorobenzene | 502.2 | PID | 2.654 | 5.904 | 2.063 |
| Chlorobenzene | 524.2 | | 0.880 | 0.643 | 4.864 |
| Chloroethane | 502.2 | ELCD | 0.051 | 0.041 | 23.132 |
| Chloroethane | 524.2 | | 0.619 | 0.637 | 5.069 |
| Chloroform | 502.2 | ELCD | 9.595 | 17.716 | 0.722 |
| Chloroform | 524.2 | | 0.745 | 0.692 | 5.610 |
| Chloromethane | 502.2 | ELCD | 1.292 | 3.452 | 2.934 |
| Chloromethane | 524.2 | | 0.803 | 1.469 | 4.798 |
| Chromium | 1620 | | 0.819 | 1.246 | 1.545 |
| Chromium | 200.8 | | 0.847 | 1.713 | 5.391 |
| Cis-1,2-dce+2,2-dcp | 502.2 | ELCD | 1.334 | 2.619 | 3.344 |
| Cis-1,2-dichloroethene | 524.2 | | 0.826 | -0.342 | 5.104 |

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|---------------------------------|--------|-----------|-----------------|-----------------|----------------|
| Cis-1,3-dichloropropene | 502.2 | ELCD | 2.183 | 4.379 | 8.267 |
| Cis-1,3-dichloropropene | 502.2 | PID | 0.839 | 1.742 | 1.367 |
| Cis-1,3-dichloropropene | 524.2 | | 0.616 | 0.038 | 4.116 |
| Cobalt | 1620 | | 0.409 | 0.824 | 1.585 |
| Cobalt | 200.8 | | 0.790 | -64.641 | 8.058 |
| Copper | 1620 | | 0.825 | 1.739 | 3.362 |
| Copper | 200.8 | | 0.716 | 1.429 | 20.667 |
| Dibromochloromethane | 502.2 | ELCD | 2.077 | 4.211 | 45.610 |
| Dibromochloromethane | 524.2 | | 0.600 | 0.208 | 4.953 |
| Dibromomethane | 502.2 | ELCD | 7.153 | 14.604 | 67.146 |
| Dibromomethane | 524.2 | | 0.806 | 1.145 | 3.356 |
| Dichlorodifluoromethane | 502.2 | ELCD | 163.453 | 192.935 | 9.771 |
| Dichlorodifluoromethane | 524.2 | | 0.651 | 0.551 | 5.027 |
| Diethyl Ether | 524.2 | | 0.952 | 1.412 | 2.837 |
| Ethyl Methacrylate | 524.2 | | 0.681 | 0.401 | 5.391 |
| Ethylbenzene | 502.2 | PID | 2.459 | 5.663 | 3.496 |
| Ethylbenzene | 524.2 | | 0.833 | -0.475 | 4.992 |
| Hardness | 130.2 | | 0.669 | 1.391 | 2.598 |
| Hexachlorobutadiene | 502.2 | ELCD | 5.586 | 11.673 | 2.097 |
| Hexachlorobutadiene | 524.2 | | 0.519 | -0.293 | 3.866 |
| Hexachloroethane | 524.2 | | 0.883 | 0.782 | 4.186 |
| Hexachlorobutadiene+naphthalene | 502.2 | PID | 1.423 | 3.208 | 0.920 |
| Iron | 1620 | | 2.991 | 5.223 | 3.824 |
| Isopropylbenzene | 502.2 | PID | 2.530 | 5.930 | 2.921 |
| Isopropylbenzene | 524.2 | | 0.920 | 1.099 | 10.555 |
| Lead | 1620 | | 0.720 | 1.193 | 1.407 |
| Lead | 200.8 | | 0.093 | 0.183 | 0.301 |
| M+p Xylene | 502.2 | PID | 1.093 | 2.496 | 1.281 |
| M+p Xylene | 524.2 | | 0.588 | 0.449 | 9.870 |
| Magnesium | 1620 | | 0.861 | 1.702 | 0.967 |
| Manganese | 1620 | | 0.158 | 0.378 | 0.953 |
| Manganese | 200.8 | | 0.974 | 1.587 | 3.450 |
| Mercury | 200.8 | | 0.799 | -4.048 | 14.165 |
| Methacrylonitrile | 524.2 | | 0.642 | 1.034 | 1.808 |

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|-------------------------|--------|-----------|-----------------|-----------------|----------------|
| Methyl Iodide | 524.2 | | 0.924 | -0.257 | 6.826 |
| Methyl Tert-butyl Ether | 524.2 | | 0.630 | -1.011 | 7.514 |
| Methylacrylate | 524.2 | | 0.918 | 1.618 | 2.495 |
| Methylene Chloride | 502.2 | ELCD | 14.312 | 39.130 | 21.261 |
| Methylene Chloride | 524.2 | | 0.875 | 1.196 | 3.367 |
| Methylmethacrylate | 524.2 | | 0.376 | 0.535 | 2.153 |
| Molybdenum | 1620 | | 0.698 | 1.543 | 1.188 |
| Molybdenum | 200.8 | | 0.777 | 0.013 | 68.949 |
| N-butylbenzene | 502.2 | PID | 2.322 | 5.050 | 4.676 |
| N-butylbenzene | 524.2 | | 0.890 | 1.694 | 8.614 |
| N-propylbenzene | 502.2 | PID | 14.846 | 33.306 | 2.221 |
| N-propylbenzene | 524.2 | | 0.676 | -1.102 | 6.104 |
| Naphthalene | 524.2 | | 0.821 | 0.983 | 3.622 |
| Nickel | 1620 | | 0.656 | 1.271 | 1.176 |
| Nickel | 200.8 | | 0.397 | 0.736 | 0.524 |
| o-xylene | 524.2 | | 0.802 | -1.396 | 8.734 |
| o-xylene+styrene | 502.2 | PID | 2.012 | 4.469 | 1.893 |
| P-isopropyltol+1,4-dcb | 502.2 | PID | 2.090 | 4.573 | 2.107 |
| Pentachloroethane | 524.2 | | 0.035 | -0.153 | 0.610 |
| Sec-butylbenzene | 502.2 | PID | 1.039 | 2.402 | 1.429 |
| Sec-butylbenzene | 524.2 | | 0.762 | -0.517 | 9.507 |
| Selenium | 1620 | | 0.729 | 1.759 | 2.256 |
| Selenium | 200.8 | | 0.815 | 1.577 | 2.139 |
| Silver | 1620 | | 0.731 | 1.324 | 2.082 |
| Silver | 200.8 | | 0.441 | 0.947 | 2.581 |
| Sodium | 1620 | | 0.713 | 1.404 | 1.913 |
| Styrene | 524.2 | | 0.797 | 0.815 | 8.467 |
| Tert-butylbenzene | 502.2 | PID | 2.024 | 4.819 | 2.575 |
| Tert-butylbenzene | 524.2 | | 0.535 | 1.065 | 7.707 |
| Tetrachloroethene | 502.2 | ELCD | 11.204 | 23.979 | 2.831 |
| Tetrachloroethene | 502.2 | PID | 5.175 | 12.216 | 2.540 |
| Tetrachloroethene | 524.2 | | 0.997 | 0.688 | 4.482 |
| Thallium | 1620 | | 1.273 | 2.747 | 2.361 |
| Thallium | 200.8 | | 1.038 | 1.964 | 3.125 |

Table 3. Ratios of Detection Limits to the EPA/ACS DL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO CRV/ MDL | ISO MDV/ MDL | SL-IDE/ MDL |
|-----------------------------|--------|-----------|-----------------|-----------------|----------------|
| Thorium | 200.8 | | 0.836 | -6.702 | 4.345 |
| Tin | 1620 | | 0.550 | 0.856 | 0.327 |
| Titanium | 1620 | | 0.932 | 1.685 | 1.097 |
| Toluene | 502.2 | PID | 0.923 | 2.068 | 0.873 |
| Toluene | 524.2 | | 0.280 | -0.100 | 6.591 |
| Total Phosphorus | 365.2 | | 0.777 | 1.572 | 2.202 |
| Total Suspended Solids | 160.2 | | 0.810 | 1.662 | 2.459 |
| Trans-1,2-dichloroethene | 502.2 | ELCD | 4.267 | 9.382 | 1.603 |
| Trans-1,2-dichloroethene | 524.2 | | 0.834 | -0.210 | 6.695 |
| Trans-1,3-dichloropropene | 502.2 | ELCD | 1.098 | 2.183 | 6.938 |
| Trans-1,3-dichloropropene | 502.2 | PID | 0.641 | 1.358 | 1.454 |
| Trans-1,3-dichloropropene | 524.2 | | 0.486 | -0.008 | 4.018 |
| Trans-1,4-dichloro-2-butene | 524.2 | | 0.680 | 1.150 | 2.310 |
| Trichloroethene | 502.2 | ELCD | 1.157 | 2.435 | 4.169 |
| Trichloroethene | 502.2 | PID | 1.592 | 3.635 | 3.539 |
| Trichloroethene | 524.2 | | 0.952 | 1.009 | 4.690 |
| Trichlorofluoromethane | 502.2 | ELCD | 0.114 | 0.256 | 18.504 |
| Trichlorofluoromethane | 524.2 | | 0.858 | 0.527 | 3.516 |
| Uranium | 200.8 | | 0.495 | 0.757 | 2.519 |
| Vanadium | 1620 | | 0.573 | 1.138 | 1.370 |
| Vanadium | 200.8 | | 0.923 | 1.791 | 1.522 |
| Vinyl Chloride | 502.2 | ELCD | 0.144 | 0.268 | 13.063 |
| Vinyl Chloride | 524.2 | | 0.719 | 0.004 | 6.845 |
| WAD Cyanide | 1677 | | 0.296 | 0.558 | 1.175 |
| Xylene (Total) | 524.2 | | 0.575 | 0.850 | 12.430 |
| Yttrium | 1620 | | 0.712 | 1.310 | 1.622 |
| Zinc | 1620 | | 0.886 | 1.423 | 1.700 |
| Zinc | 200.8 | | 0.511 | 0.895 | 1.663 |

Note: ELCD or PID in the Procedure column indicates the photo-ionization detector (PID) or electrolytic conductivity detector (ELCD) in EPA Method 502.2

Summary Statistics for Table 3

| | ISO CRV/ EPA/ACS DL | ISO MDV/ EPA/ACS DL | SL-IDE/ EPA/ACS DL |
|-----------------|------------------------|------------------------|---------------------------|
| Minimum | 0.01 | -64.64 | 0.10 |
| 25th percentile | 0.61 | 0.52 | 1.60 |
| Median | 0.80 | 1.17 | 2.86 |
| 75th percentile | 1.08 | 2.17 | 5.41 |
| Maximum | 163.45 | 192.93 | 242.50 |
| | | | |
| | Number of analytes | Percent of analytes | p-value for percent=50 |
| CRV>DL | 52 | 26.26% | <0.0001 |
| MDV>DL | 113 | 57.07% | 0.055 |
| SL-IDE>DL | 181 | 91.41% | <0.0001 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|---------------------------|--------|-----------|----------------|-------------------|----------------|
| 1,1,1,2-tetrachloroethane | 502.2 | ELCD | 0.2 | 0.152 | 0.030 |
| 1,1,1,2-tetrachloroethane | 524.2 | | 0.2 | 0.183 | 0.181 |
| 1,1,1-trichloroethane | 502.2 | ELCD | 0.1 | 0.044 | 0.830 |
| 1,1,1-trichloroethane | 524.2 | | 0.5 | 0.102 | 0.240 |
| 1,1,2,2-tce+1,2,3-tcp | 502.2 | ELCD | 0.5 | 1.110 | 5.514 |
| 1,1,2,2-tetrachloroethane | 524.2 | | 0.5 | 0.596 | 0.569 |
| 1,1,2-trichloroethane | 502.2 | ELCD | 0.1 | 0.289 | 0.060 |
| 1,1,2-trichloroethane | 524.2 | | 0.5 | 0.212 | 0.290 |
| 1,1-dichloroethane | 502.2 | ELCD | 0.05 | 0.047 | 0.527 |
| 1,1-dichloroethane | 524.2 | | 0.2 | 0.099 | 0.115 |
| 1,1-dichloroethene | 502.2 | ELCD | 0.1 | 0.096 | 3.796 |
| 1,1-dichloroethene | 524.2 | | 0.2 | 0.159 | 0.129 |
| 1,1-dichloropropanone | 524.2 | | 20 | 15.396 | 12.705 |
| 1,1-dichloropropene | 524.2 | | 0.5 | 0.057 | 0.180 |
| 1,2,3-trichlorobenzene | 502.2 | ELCD | 0.2 | 1.750 | 0.851 |
| 1,2,3-trichlorobenzene | 502.2 | PID | 0.2 | 1.818 | 0.248 |
| 1,2,3-trichlorobenzene | 524.2 | | 0.2 | 0.192 | 0.216 |
| 1,2,3-trichloropropane | 524.2 | | 20 | 0.267 | 11.316 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|-----------------------------|--------|-----------|----------------|-------------------|--------------------|
| 1,2,4-trichlorobenzene | 502.2 | ELCD | 0.1 | 1.150 | 0.401 |
| 1,2,4-trichlorobenzene | 502.2 | PID | 0.2 | 1.371 | 0.439 |
| 1,2,4-trichlorobenzene | 524.2 | | 0.2 | 0.231 | 0.141 |
| 1,2,4-trimethylbenzene | 502.2 | PID | 0.5 | 0.485 | 0.653 |
| 1,2,4-trimethylbenzene | 524.2 | | 0.1 | 0.050 | 20.896 |
| 1,2-dibromo-3-chloropropane | 524.2 | | 10 | 1.843 | 71.182 |
| 1,2-dibromoethane | 502.2 | ELCD | 0.5 | 0.156 | 0.592 |
| 1,2-dibromoethane | 524.2 | | 0.5 | 0.560 | 0.417 |
| 1,2-dichlorobenzene | 502.2 | ELCD | 0.1 | 0.418 | 0.183 |
| 1,2-dichlorobenzene | 502.2 | PID | 0.1 | 0.139 | 0.346 |
| 1,2-dichlorobenzene | 524.2 | | 0.1 | 0.101 | 0.085 |
| 1,2-dichloroethane | 502.2 | ELCD | 0.1 | 0.089 | 0.065 |
| 1,2-dichloroethane | 524.2 | | 0.5 | 0.122 | 0.222 |
| 1,2-dichloropropane | 502.2 | ELCD | 0.1 | 1.138 | 0.102 |
| 1,2-dichloropropane | 524.2 | | 0.5 | 0.148 | 0.196 |
| 1,3,5-tmb+4-chlorotoluene | 502.2 | PID | 0.2 | 1.275 | 0.189 |
| 1,3,5-trimethylbenzene | 524.2 | | 0.1 | 0.044 | 23.744 |
| 1,3-dichlorobenzene | 502.2 | ELCD | 0.1 | 0.301 | 0.936 |
| 1,3-dichlorobenzene | 502.2 | PID | 0.2 | 0.840 | 0.465 |
| 1,3-dichlorobenzene | 524.2 | | 0.1 | 0.080 | 0.076 |
| 1,3-dichloropropane | 502.2 | ELCD | 0.1 | 0.356 | 0.054 |
| 1,3-dichloropropane | 524.2 | | 0.2 | 0.114 | 0.139 |
| 1,4-dichlorobenzene | 502.2 | ELCD | 0.1 | 0.571 | 0.101 |
| 1,4-dichlorobenzene | 524.2 | | 0.1 | 0.069 | 0.078 |
| 1-chlorobutane | 524.2 | | 0.1 | 0.082 | 29.943 |
| 2,2-dichloropropane | 524.2 | | 10 | 0.568 | 38.009 |
| 2-butanone | 524.2 | | 2 | 1.416 | 0.893 |
| 2-chlorotoluene | 502.2 | ELCD | 0.5 | 0.881 | 0.493 |
| 2-chlorotoluene | 502.2 | PID | 1 | 2.877 | 0.849 |
| 2-chlorotoluene | 524.2 | | 0.1 | 0.046 | 0.053 |
| 2-hexanone | 524.2 | | 10 | 0.669 | 0.442 |
| 2-nitropropane | 524.2 | | 10 | 1.280 | 0.590 |
| 4-chlorotoluene | 502.2 | ELCD | 0.5 | 0.668 | 0.142 ¹ |
| 4-chlorotoluene | 524.2 | | 0.1 | 0.037 | 23.810 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|----------------------------------|--------|-----------|----------------|------------------------|----------------|
| 4-isopropyltoluene | 524.2 | | 0.1 | 0.043 | 0.016 |
| 4-methyl-2-pentanone | 524.2 | | 5 | 2.065 | 1.785 |
| Acetone | 524.2 | | 2 | 2.115 | 2.741 |
| Acrylonitrile | 524.2 | | 5 | 1.816 | 28.056 |
| Allyl Chloride | 524.2 | | 0.2 | 0.129 | 29.674 |
| Aluminum | 1620 | | 100 | 76.242 | 464.069 |
| Aluminum | 200.8 | | 50 | 9.418 | 29.684 |
| Ammonia as Nitrogen ² | 350.3 | | 0.05 | 0.037 | 0.035 |
| Antimony | 1620 | | 5 | 4.784 | 9.551 |
| Antimony | 200.8 | | 0.5 | 0.017 | 0.034 |
| Arsenic | 1620 | | 5 | 3.684 | 3.097 |
| Arsenic | 200.8 | | 1 | 0.720 | 0.798 |
| Barium | 1620 | | 5 | 4.722 | 4.118 |
| Barium | 200.8 | | 0.1 | 0.161 | 0.211 |
| Benzene | 502.2 | PID | 0.1 | 0.273 | 0.182 |
| Benzene | 524.2 | | 0.05 | 0.075 | 0.044 |
| Beryllium | 1620 | | 2 | 1.055 | 0.980 |
| Beryllium | 200.8 | | 0.02 | 0.018 | 0.044 |
| Boron | 1620 | | 100 | 46.040 | 51.134 |
| Bromobenzene | 502.2 | ELCD | 0.5 | 0.593 | 3.529 |
| Bromobenzene | 502.2 | PID | 0.05 | 1.767 | 0.100 |
| Bromobenzene | 524.2 | | 0.2 | 0.167 | 0.140 |
| Bromochloromethane | 502.2 | ELCD | 0.1 | 0.090 | 1.598 |
| Bromochloromethane | 524.2 | | 0.5 | 0.549 | 0.368 |
| Bromodichloromethane | 502.2 | ELCD | 0.05 | 0.091 | 0.424 |
| Bromodichloromethane | 524.2 | | 0.2 | 0.135 | 0.128 |
| Bromoform | 502.2 | ELCD | 0.2 | 0.056 | 3.393 |
| Bromoform | 524.2 | | 0.5 | 0.287 | 0.482 |
| Bromomethane | 502.2 | ELCD | 1 | undefined ³ | 16.351 |
| Bromomethane | 524.2 | | 0.2 | 0.252 | 0.226 |
| Cadmium | 1620 | | 0.5 | 0.346 | 0.410 |
| Cadmium | 200.8 | | 0.02 | 0.046 | 0.063 |
| Calcium | 1620 | | 100 | 186.530 | 99.975 |
| Carbon Disulfide | 524.2 | | 0.1 | 0.077 | 0.101 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|-------------------------|--------|-----------|----------------|------------------------|------------------------|
| Carbon Tetrachloride | 524.2 | | 0.2 | 0.127 | 0.140 |
| Carbontet+1,1-dcp | 502.2 | ELCD | 0.2 | 0.173 | 0.069 |
| Chloroacetonitrile | 524.2 | | 10 | 4.169 | 3.310 |
| Chlorobenzene | 502.2 | ELCD | 0.1 | 0.092 | 1.766 |
| Chlorobenzene | 502.2 | PID | 0.1 | 0.504 | 0.119 |
| Chlorobenzene | 524.2 | | 0.1 | 0.108 | 0.059 |
| Chloroethane | 502.2 | ELCD | 0.5 | 0.037 | 5.826 |
| Chloroethane | 524.2 | | 0.5 | 0.185 | 0.255 |
| Chloroform | 502.2 | ELCD | 0.2 | 2.217 | 0.025 |
| Chloroform | 524.2 | | 0.2 | 0.138 | 0.121 |
| Chloromethane | 502.2 | ELCD | 0.2 | 0.709 | 1.734 |
| Chloromethane | 524.2 | | 0.2 | 0.181 | 0.141 |
| Chromium | 1620 | | 1 | 0.993 | 1.259 |
| Chromium | 200.8 | | 0.2 | 0.331 | 1.028 |
| Cis-1,2-dce+2,2-dcp | 502.2 | ELCD | 0.05 | 0.097 | 0.039 |
| Cis-1,2-dichloroethene | 524.2 | | 0.5 | 0.154 | 0.144 |
| Cis-1,3-dichloropropene | 502.2 | ELCD | 0.1 | 0.090 | 0.415 |
| Cis-1,3-dichloropropene | 502.2 | PID | 0.2 | 0.254 | 0.017 ¹ |
| Cis-1,3-dichloropropene | 524.2 | | 0.2 | 0.116 | 0.141 |
| Cobalt | 1620 | | 50 | 20.916 | 40.837 |
| Cobalt | 200.8 | | 0.005 | undefined ³ | undefined ⁴ |
| Copper | 1620 | | 20 | 27.513 | 47.509 |
| Copper | 200.8 | | 0.1 | 0.142 | 1.825 |
| Dibromochloromethane | 502.2 | ELCD | 0.1 | 0.106 | 1.252 |
| Dibromochloromethane | 524.2 | | 0.5 | 0.149 | 0.288 |
| Dibromomethane | 502.2 | ELCD | 0.1 | 0.257 | 1.395 |
| Dibromomethane | 524.2 | | 0.5 | 0.400 | 0.460 |
| Dichlorodifluoromethane | 502.2 | ELCD | 0.1 | 5.759 | 1.091 |
| Dichlorodifluoromethane | 524.2 | | 0.5 | 0.289 | 0.480 |
| Diethyl Ether | 524.2 | | 0.5 | 0.563 | 0.404 |
| Ethyl Methacrylate | 524.2 | | 0.5 | 0.139 | 0.183 |
| Ethylbenzene | 502.2 | PID | 0.1 | 0.341 | 0.157 |
| Ethylbenzene | 524.2 | | 0.2 | 0.123 | 0.077 |
| Hardness ² | 130.2 | | 2 | 2.973 | 5.465 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|---------------------------------|--------|-----------|----------------|------------------------|----------------|
| Hexachlorobutadiene | 502.2 | ELCD | 0.2 | 1.466 | 0.243 |
| Hexachlorobutadiene | 524.2 | | 0.2 | 0.160 | 0.228 |
| Hexachloroethane | 524.2 | | 0.2 | 0.232 | 0.167 |
| Hexachlorobutadiene+naphthalene | 502.2 | PID | 2 | 6.108 | 1.542 |
| Iron | 1620 | | 200 | 1490.589 | 996.565 |
| Isopropylbenzene | 502.2 | PID | 0.1 | 0.337 | 0.129 |
| Isopropylbenzene | 524.2 | | 0.1 | 0.056 | 25.592 |
| Lead | 1620 | | 5 | 5.062 | 5.698 |
| Lead | 200.8 | | 2 | 0.318 | 0.685 |
| M+p Xylene | 502.2 | PID | 0.2 | 0.652 | 0.222 |
| M+p Xylene | 524.2 | | 0.1 | 0.042 | 24.651 |
| Magnesium | 1620 | | 500 | 454.043 | 267.199 |
| Manganese | 1620 | | 20 | 7.948 | 15.264 |
| Manganese | 200.8 | | 0.1 | 0.133 | 0.245 |
| Mercury | 200.8 | | 0.02 | 0.056 | 0.039 |
| Methacrylonitrile | 524.2 | | 2 | 1.065 | 19.062 |
| Methyl Iodide | 524.2 | | 0.2 | 0.108 | 0.083 |
| Methyl Tert-butyl Ether | 524.2 | | 0.5 | 0.073 | 0.122 |
| Methylacrylate | 524.2 | | 1 | 0.966 | 0.727 |
| Methylene Chloride | 502.2 | ELCD | 1 | undefined ³ | 6.033 |
| Methylene Chloride | 524.2 | | 0.2 | 0.354 | 0.433 |
| Methylmethacrylate | 524.2 | | 1 | 0.381 | 20.773 |
| Molybdenum | 1620 | | 10 | 9.752 | 7.597 |
| Molybdenum | 200.8 | | 0.01 | 0.052 | 0.608 |
| N-butylbenzene | 502.2 | PID | 0.1 | 0.429 | 0.745 |
| N-butylbenzene | 524.2 | | 0.1 | 0.077 | 0.067 |
| N-propylbenzene | 502.2 | PID | 0.2 | 3.869 | 0.186 |
| N-propylbenzene | 524.2 | | 0.2 | 0.110 | 29.878 |
| Naphthalene | 524.2 | | 0.2 | 0.184 | 0.108 |
| Nickel | 1620 | | 100 | 66.486 | 67.206 |
| Nickel | 200.8 | | 0.5 | 0.287 | 0.183 |
| o-xylene | 524.2 | | 0.2 | 0.062 | 0.040 |
| o-xylene+styrene | 502.2 | PID | 0.2 | 0.746 | 0.181 |
| P-isopropylol+1,4-dcb | 502.2 | PID | 0.2 | 0.956 | 0.456 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|-------------------------------------|--------|-----------|----------------|-------------------|----------------|
| Pentachloroethane | 524.2 | | 2 | 0.086 | 0.551 |
| Sec-butylbenzene | 502.2 | PID | 0.2 | 0.377 | 0.157 |
| Sec-butylbenzene | 524.2 | | 0.1 | 0.063 | 0.047 |
| Selenium | 1620 | | 2 | 3.859 | 5.235 |
| Selenium | 200.8 | | 1 | 0.805 | 1.045 |
| Silver | 1620 | | 20 | 16.734 | 25.842 |
| Silver | 200.8 | | 0.02 | 0.011 | 0.056 |
| Sodium | 1620 | | 200 | 251.546 | 337.755 |
| Styrene | 524.2 | | 0.1 | 0.054 | 0.041 |
| Tert-butylbenzene | 502.2 | PID | 0.1 | 0.391 | 0.203 |
| Tert-butylbenzene | 524.2 | | 0.1 | 0.063 | 0.073 |
| Tetrachloroethene | 502.2 | ELCD | 0.1 | 1.226 | 0.122 |
| Tetrachloroethene | 502.2 | PID | 0.2 | 2.084 | 0.750 |
| Tetrachloroethene | 524.2 | | 0.2 | 0.377 | 30.554 |
| Thallium | 1620 | | 2 | 3.748 | 2.799 |
| Thallium | 200.8 | | 0.001 | 0.002 | 0.003 |
| Thorium | 200.8 | | 0.002 | 0.005 | 0.004 |
| Tin | 1620 | | 10 | 9.237 | 9.406 |
| Titanium | 1620 | | 20 | 20.807 | 14.236 |
| Toluene | 502.2 | PID | 0.2 | 0.409 | 0.194 |
| Toluene | 524.2 | | 0.2 | 0.028 | 0.046 |
| Total Phosphorus ² | 365.2 | | 0.02 | 0.024 | 0.030 |
| Total Suspended Solids ² | 160.2 | | 5 | 5.011 | 6.729 |
| Trans-1,2-dichloroethene | 502.2 | ELCD | 0.2 | 1.060 | 0.191 |
| Trans-1,2-dichloroethene | 524.2 | | 0.2 | 0.140 | 0.153 |
| Trans-1,3-dichloropropene | 502.2 | ELCD | 0.1 | 0.073 | 0.729 |
| Trans-1,3-dichloropropene | 502.2 | PID | 0.2 | 0.205 | 0.175 |
| Trans-1,3-dichloropropene | 524.2 | | 0.5 | 0.121 | 0.218 |
| Trans-1,4-dichloro-2-butene | 524.2 | | 2 | 1.801 | 30.108 |
| Trichloroethene | 502.2 | ELCD | 0.1 | 0.081 | 3.169 |
| Trichloroethene | 502.2 | PID | 0.1 | 0.260 | 0.401 |
| Trichloroethene | 524.2 | | 0.2 | 0.283 | 0.167 |
| Trichlorofluoromethane | 502.2 | ELCD | 0.5 | 0.085 | 4.662 |
| Trichlorofluoromethane | 524.2 | | 0.5 | 0.279 | 42.490 |

Table 4. Comparison Quantitation Limits for the Episode 6000 Data Set
(µg/L except where footnoted)

| Analyte | Method | Procedure | EPA/ ACS QL | ISO/ IUPAC LOQ | ASTM SL-IQE |
|----------------|--------|-----------|----------------|-------------------|----------------|
| Uranium | 200.8 | | 0.001 | 0.001 | 0.002 |
| Vanadium | 1620 | | 20 | 21.586 | 24.338 |
| Vanadium | 200.8 | | 2 | 2.627 | 1.933 |
| Vinyl Chloride | 502.2 | ELCD | 1 | 0.270 | 8.234 |
| Vinyl Chloride | 524.2 | | 0.2 | 0.139 | 0.219 |
| WAD Cyanide | 1677 | | 2 | 0.852 | 1.624 |
| Xylene (Total) | 524.2 | | 0.1 | 0.027 | 23.520 |
| Yttrium | 1620 | | 5 | 6.571 | 8.962 |
| Zinc | 1620 | | 10 | 9.575 | 10.452 |
| Zinc | 200.8 | | 2 | 2.147 | 7.024 |

¹ IQE 10% undefined, IQE 20% reported² Results reported as mg/L³ No LOQ could be calculated due to a square root of a negative number in the formula⁴ IQE 10%, IQE 20% and IQE 30% all negative based on chosen model (linear)Note: ELCD or PID in the Procedure column indicates the photo-ionization detector (PID) or electrolytic conductivity detector (ELCD) in EPA Method 502.2**Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set**

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|---------------------------|--------|-----------|---------------|-----------|
| 1,1,1,2-tetrachloroethane | 502.2 | ELCD | 0.759 | 0.152 |
| 1,1,1,2-tetrachloroethane | 524.2 | | 0.915 | 0.907 |
| 1,1,1-trichloroethane | 502.2 | ELCD | 0.445 | 8.298 |
| 1,1,1-trichloroethane | 524.2 | | 0.203 | 0.479 |
| 1,1,2,2-tce+1,2,3-tcp | 502.2 | ELCD | 2.220 | 11.027 |
| 1,1,2,2-tetrachloroethane | 524.2 | | 1.192 | 1.138 |
| 1,1,2-trichloroethane | 502.2 | ELCD | 2.886 | 0.603 |
| 1,1,2-trichloroethane | 524.2 | | 0.424 | 0.579 |
| 1,1-dichloroethane | 502.2 | ELCD | 0.949 | 10.531 |
| 1,1-dichloroethane | 524.2 | | 0.496 | 0.573 |
| 1,1-dichloroethene | 502.2 | ELCD | 0.960 | 37.963 |
| 1,1-dichloroethene | 524.2 | | 0.795 | 0.644 |
| 1,1-dichloropropanone | 524.2 | | 0.770 | 0.635 |
| 1,1-dichloropropene | 524.2 | | 0.114 | 0.360 |
| 1,2,3-trichlorobenzene | 502.2 | ELCD | 8.751 | 4.257 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|-----------------------------|--------|-----------|------------|-----------|
| 1,2,3-trichlorobenzene | 502.2 | PID | 9.092 | 1.239 |
| 1,2,3-trichlorobenzene | 524.2 | | 0.959 | 1.080 |
| 1,2,3-trichloropropane | 524.2 | | 0.013 | 0.566 |
| 1,2,4-trichlorobenzene | 502.2 | ELCD | 11.502 | 4.010 |
| 1,2,4-trichlorobenzene | 502.2 | PID | 6.853 | 2.197 |
| 1,2,4-trichlorobenzene | 524.2 | | 1.157 | 0.703 |
| 1,2,4-trimethylbenzene | 502.2 | PID | 0.970 | 1.306 |
| 1,2,4-trimethylbenzene | 524.2 | | 0.503 | 208.956 |
| 1,2-dibromo-3-chloropropane | 524.2 | | 0.184 | 7.118 |
| 1,2-dibromoethane | 502.2 | ELCD | 0.311 | 1.185 |
| 1,2-dibromoethane | 524.2 | | 1.119 | 0.834 |
| 1,2-dichlorobenzene | 502.2 | ELCD | 4.178 | 1.833 |
| 1,2-dichlorobenzene | 502.2 | PID | 1.393 | 3.455 |
| 1,2-dichlorobenzene | 524.2 | | 1.006 | 0.848 |
| 1,2-dichloroethane | 502.2 | ELCD | 0.891 | 0.650 |
| 1,2-dichloroethane | 524.2 | | 0.244 | 0.443 |
| 1,2-dichloropropane | 502.2 | ELCD | 11.378 | 1.023 |
| 1,2-dichloropropane | 524.2 | | 0.296 | 0.393 |
| 1,3,5-tmb+4-chlorotoluene | 502.2 | PID | 6.373 | 0.946 |
| 1,3,5-trimethylbenzene | 524.2 | | 0.436 | 237.442 |
| 1,3-dichlorobenzene | 502.2 | ELCD | 3.008 | 9.356 |
| 1,3-dichlorobenzene | 502.2 | PID | 4.201 | 2.326 |
| 1,3-dichlorobenzene | 524.2 | | 0.800 | 0.760 |
| 1,3-dichloropropane | 502.2 | ELCD | 3.560 | 0.539 |
| 1,3-dichloropropane | 524.2 | | 0.569 | 0.695 |
| 1,4-dichlorobenzene | 502.2 | ELCD | 5.714 | 1.010 |
| 1,4-dichlorobenzene | 524.2 | | 0.686 | 0.784 |
| 1-chlorobutane | 524.2 | | 0.823 | 299.428 |
| 2,2-dichloropropane | 524.2 | | 0.057 | 3.801 |
| 2-butanone | 524.2 | | 0.708 | 0.446 |
| 2-chlorotoluene | 502.2 | ELCD | 1.763 | 0.986 |
| 2-chlorotoluene | 502.2 | PID | 2.877 | 0.849 |
| 2-chlorotoluene | 524.2 | | 0.463 | 0.533 |
| 2-hexanone | 524.2 | | 0.067 | 0.044 |
| 2-nitropropane | 524.2 | | 0.128 | 0.059 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|----------------------|--------|-----------|------------|-----------|
| 4-chlorotoluene | 502.2 | ELCD | 1.335 | 0.284 |
| 4-chlorotoluene | 524.2 | | 0.373 | 238.097 |
| 4-isopropyltoluene | 524.2 | | 0.428 | 0.163 |
| 4-methyl-2-pentanone | 524.2 | | 0.413 | 0.357 |
| Acetone | 524.2 | | 1.057 | 1.371 |
| Acrylonitrile | 524.2 | | 0.363 | 5.611 |
| Allyl chloride | 524.2 | | 0.646 | 148.372 |
| Aluminum | 1620 | | 0.762 | 4.641 |
| Aluminum | 200.8 | ICP/MS | 0.188 | 0.594 |
| Ammonia as nitrogen | 350.3 | | 0.733 | 0.709 |
| Antimony | 1620 | | 0.957 | 1.910 |
| Antimony | 200.8 | ICP/MS | 0.035 | 0.067 |
| Arsenic | 1620 | | 0.737 | 0.619 |
| Arsenic | 200.8 | ICP/MS | 0.720 | 0.798 |
| Barium | 1620 | | 0.944 | 0.824 |
| Barium | 200.8 | ICP/MS | 1.608 | 2.114 |
| Benzene | 502.2 | PID | 2.735 | 1.819 |
| Benzene | 524.2 | | 1.509 | 0.877 |
| Beryllium | 1620 | | 0.527 | 0.490 |
| Beryllium | 200.8 | ICP/MS | 0.905 | 2.201 |
| Boron | 1620 | | 0.460 | 0.511 |
| Bromobenzene | 502.2 | ELCD | 1.186 | 7.058 |
| Bromobenzene | 502.2 | PID | 35.338 | 2.007 |
| Bromobenzene | 524.2 | | 0.833 | 0.699 |
| Bromochloromethane | 502.2 | ELCD | 0.900 | 15.977 |
| Bromochloromethane | 524.2 | | 1.097 | 0.737 |
| Bromodichloromethane | 502.2 | ELCD | 1.814 | 8.484 |
| Bromodichloromethane | 524.2 | | 0.676 | 0.640 |
| Bromoform | 502.2 | ELCD | 0.282 | 16.964 |
| Bromoform | 524.2 | | 0.573 | 0.964 |
| Bromomethane | 502.2 | ELCD | N/A | 16.351 |
| Bromomethane | 524.2 | | 1.262 | 1.130 |
| Cadmium | 1620 | | 0.692 | 0.820 |
| Cadmium | 200.8 | ICP/MS | 2.311 | 3.148 |
| Calcium | 1620 | | 1.865 | 1.000 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|-------------------------|--------|-----------|------------|-----------|
| Carbon disulfide | 524.2 | | 0.768 | 1.013 |
| Carbon tetrachloride | 524.2 | | 0.635 | 0.700 |
| Carbontet+1,1-dcp | 502.2 | ELCD | 0.864 | 0.343 |
| Chloroacetonitrile | 524.2 | | 0.417 | 0.331 |
| Chlorobenzene | 502.2 | ELCD | 0.921 | 17.658 |
| Chlorobenzene | 502.2 | PID | 5.040 | 1.190 |
| Chlorobenzene | 524.2 | | 1.077 | 0.595 |
| Chloroethane | 502.2 | ELCD | 0.074 | 11.652 |
| Chloroethane | 524.2 | | 0.369 | 0.510 |
| Chloroform | 502.2 | ELCD | 11.086 | 0.126 |
| Chloroform | 524.2 | | 0.688 | 0.606 |
| Chloromethane | 502.2 | ELCD | 3.547 | 8.669 |
| Chloromethane | 524.2 | | 0.906 | 0.703 |
| Chromium | 1620 | | 0.993 | 1.259 |
| Chromium | 200.8 | ICP/MS | 1.653 | 5.142 |
| Cis-1,2-dce+2,2-dcp | 502.2 | ELCD | 1.941 | 0.780 |
| Cis-1,2-dichloroethene | 524.2 | | 0.307 | 0.288 |
| Cis-1,3-dichloropropene | 502.2 | ELCD | 0.902 | 4.151 |
| Cis-1,3-dichloropropene | 502.2 | PID | 1.271 | 0.083 |
| Cis-1,3-dichloropropene | 524.2 | | 0.582 | 0.706 |
| Cobalt | 1620 | | 0.418 | 0.817 |
| Cobalt | 200.8 | ICP/MS | N/A | N/A |
| Copper | 1620 | | 1.376 | 2.375 |
| Copper | 200.8 | ICP/MS | 1.419 | 18.250 |
| Dibromochloromethane | 502.2 | ELCD | 1.059 | 12.515 |
| Dibromochloromethane | 524.2 | | 0.299 | 0.575 |
| Dibromomethane | 502.2 | ELCD | 2.572 | 13.949 |
| Dibromomethane | 524.2 | | 0.800 | 0.920 |
| Dichlorodifluoromethane | 502.2 | ELCD | 57.590 | 10.907 |
| Dichlorodifluoromethane | 524.2 | | 0.579 | 0.959 |
| Diethyl ether | 524.2 | | 1.126 | 0.808 |
| Ethyl methacrylate | 524.2 | | 0.279 | 0.366 |
| Ethylbenzene | 502.2 | PID | 3.413 | 1.573 |
| Ethylbenzene | 524.2 | | 0.613 | 0.387 |
| Hardness | 130.2 | | 1.487 | 2.733 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|---------------------------------|--------|-----------|------------|-----------|
| Hexachlorobutadiene | 502.2 | ELCD | 7.332 | 1.215 |
| Hexachlorobutadiene | 524.2 | | 0.799 | 1.142 |
| Hexachloroethane | 524.2 | | 1.159 | 0.837 |
| Hexachlorobutadiene+naphthalene | 502.2 | PID | 3.054 | 0.771 |
| Iron | 1620 | | 7.453 | 4.983 |
| Isopropylbenzene | 502.2 | PID | 3.372 | 1.290 |
| Isopropylbenzene | 524.2 | | 0.563 | 255.921 |
| Lead | 1620 | | 1.012 | 1.140 |
| Lead | 200.8 | ICP/MS | 0.159 | 0.342 |
| M+p xylene | 502.2 | PID | 3.262 | 1.112 |
| M+p xylene | 524.2 | | 0.419 | 246.513 |
| Magnesium | 1620 | | 0.908 | 0.534 |
| Manganese | 1620 | | 0.397 | 0.763 |
| Manganese | 200.8 | ICP/MS | 1.326 | 2.451 |
| Mercury | 200.8 | ICP/MS | 2.781 | 1.933 |
| Methacrylonitrile | 524.2 | | 0.533 | 9.531 |
| Methyl iodide | 524.2 | | 0.538 | 0.417 |
| Methyl tert-butyl ether | 524.2 | | 0.146 | 0.245 |
| Methylacrylate | 524.2 | | 0.966 | 0.727 |
| Methylene chloride | 502.2 | ELCD | N/A | 6.033 |
| Methylene chloride | 524.2 | | 1.769 | 2.164 |
| Methylmethacrylate | 524.2 | | 0.381 | 20.773 |
| Molybdenum | 1620 | | 0.975 | 0.760 |
| Molybdenum | 200.8 | ICP/MS | 5.181 | 60.817 |
| N-butylbenzene | 502.2 | PID | 4.290 | 7.453 |
| N-butylbenzene | 524.2 | | 0.767 | 0.673 |
| N-propylbenzene | 502.2 | PID | 19.343 | 0.931 |
| N-propylbenzene | 524.2 | | 0.552 | 149.392 |
| Naphthalene | 524.2 | | 0.921 | 0.541 |
| Nickel | 1620 | | 0.665 | 0.672 |
| Nickel | 200.8 | ICP/MS | 0.574 | 0.365 |
| O-xylene | 524.2 | | 0.310 | 0.202 |
| O-xylene+styrene | 502.2 | PID | 3.731 | 0.905 |
| P-isoproptol+1,4-dcb | 502.2 | PID | 4.778 | 2.281 |
| Pentachloroethane | 524.2 | | 0.043 | 0.276 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|-----------------------------|--------|-----------|------------|-----------|
| Sec-butylbenzene | 502.2 | PID | 1.885 | 0.784 |
| Sec-butylbenzene | 524.2 | | 0.628 | 0.475 |
| Selenium | 1620 | | 1.929 | 2.617 |
| Selenium | 200.8 | ICP/MS | 0.805 | 1.045 |
| Silver | 1620 | | 0.837 | 1.292 |
| Silver | 200.8 | ICP/MS | 0.543 | 2.800 |
| Sodium | 1620 | | 1.258 | 1.689 |
| Styrene | 524.2 | | 0.536 | 0.407 |
| Tert-butylbenzene | 502.2 | PID | 3.911 | 2.027 |
| Tert-butylbenzene | 524.2 | | 0.634 | 0.735 |
| Tetrachloroethene | 502.2 | ELCD | 12.264 | 1.217 |
| Tetrachloroethene | 502.2 | PID | 10.419 | 3.749 |
| Tetrachloroethene | 524.2 | | 1.886 | 152.769 |
| Thallium | 1620 | | 1.874 | 1.399 |
| Thallium | 200.8 | ICP/MS | 2.085 | 3.286 |
| Thorium | 200.8 | ICP/MS | 2.648 | 2.198 |
| Tin | 1620 | | 0.924 | 0.941 |
| Titanium | 1620 | | 1.040 | 0.712 |
| Toluene | 502.2 | PID | 2.045 | 0.971 |
| Toluene | 524.2 | | 0.138 | 0.228 |
| Total phosphorus | 365.2 | | 1.188 | 1.499 |
| Total suspended solids | 160.2 | | 1.002 | 1.346 |
| Trans-1,2-dichloroethene | 502.2 | ELCD | 5.298 | 0.953 |
| Trans-1,2-dichloroethene | 524.2 | | 0.702 | 0.764 |
| Trans-1,3-dichloropropene | 502.2 | ELCD | 0.731 | 7.286 |
| Trans-1,3-dichloropropene | 502.2 | PID | 1.023 | 0.874 |
| Trans-1,3-dichloropropene | 524.2 | | 0.241 | 0.436 |
| Trans-1,4-dichloro-2-butene | 524.2 | | 0.900 | 15.054 |
| Trichloroethene | 502.2 | ELCD | 0.812 | 31.690 |
| Trichloroethene | 502.2 | PID | 2.601 | 4.010 |
| Trichloroethene | 524.2 | | 1.417 | 0.837 |
| Trichlorofluoromethane | 502.2 | ELCD | 0.171 | 9.325 |
| Trichlorofluoromethane | 524.2 | | 0.558 | 84.980 |
| Uranium | 200.8 | ICP/MS | 0.721 | 1.853 |
| Vanadium | 1620 | | 1.079 | 1.217 |

Table 5. Ratios of Quantitation Limits to the EPA/ACS QL for the Episode 6000 Data Set

| Analyte | Method | Procedure | ISO LOQ/ML | SL-IQE/ML |
|----------------|--------|-----------|------------|-----------|
| Vanadium | 200.8 | ICP/MS | 1.314 | 0.966 |
| Vinyl chloride | 502.2 | ELCD | 0.270 | 8.234 |
| Vinyl chloride | 524.2 | | 0.697 | 1.097 |
| Wad cyanide | 1677 | WADCN | 0.426 | 0.812 |
| Xylene (total) | 524.2 | | 0.266 | 235.197 |
| Yttrium | 1620 | | 1.314 | 1.792 |
| Zinc | 1620 | | 0.957 | 1.045 |
| Zinc | 200.8 | ICP/MS | 1.073 | 3.512 |

Note: ELCD or PID in the Procedure column indicates the photo-ionization detector (PID) or electrolytic conductivity detector (ELCD) in EPA Method 502.2

Summary Statistics for Table 5

| | ISO LOQ/QL | SL-IQE/QL | |
|-----------------|--------------------|---------------------|------------------------|
| Minimum | 0.013 | 0.044 | |
| 25th percentile | 0.547 | 0.672 | |
| Median | 0.915 | 1.023 | |
| 75th percentile | 1.792 | 3.512 | |
| Maximum | 57.590 | 299.428 | |
| | | | |
| | Number of analytes | Percent of analytes | p-value for percent=50 |
| LOQ>QL | 84 | 43.08% | 0.062 |
| SL-IQE>QL | 101 | 51.27% | 0.776 |

Table 6. Frequency Comparisons for Lowest and Highest Detection and Quantitation Limits for the Episode 6000 Data Set

| Detection limits | | | | |
|----------------------------|-------------------|----------------------|----------------|--------------------|
| Rank | EPA/ACS DL | ISO CRV | ISO MDV | ASTM SL-IDE |
| 1 (lowest) | 30.3% | 49.5% | 17.7% | 2.5% |
| 2 | 34.3% | 33.3% | 16.2% | 16.2% |
| 3 | 30.8% | 17.2% | 33.3% | 18.7% |
| 4 (highest) | 4.5% | 0% | 32.8% | 62.6% |
| Quantitation limits | | | | |
| Rank | EPA/ACS QL | No equivalent | ISO LOQ | ASTM SL-IQE |
| 1 (lowest) | 50.5% | - | 25.8% | 22.2% |
| 2 | 33.8% | - | 26.8% | 37.9% |
| 3 (highest) | 14.1% | - | 46.0% | 38.4% |

Table 7. Detection and Quantitation Limits for EPA Methods 1631 and 1638 as Computed by EPA and by EPRI (ng/L)

| Element¹ | Ambient WQC² | Detection limits | | | Quantitation limits | | |
|----------------------------|--------------------------------|-------------------------|------------------------|-------------|----------------------------|------------------------|-------------|
| | | MDL in Method | IDE computed by | | ML in Method | IQE computed by | |
| | | | EPA | EPRI | | EPA | EPRI |
| Antimony | 14000 | 9.7 | 140 | 110 | 20 | 270 | 270 |
| Cadmium | 370 | 25 | 150 | 150 | 100 | 540 | 380 |
| Copper | 2400 | 87 | 780 | 770 | 200 | 3800 | 3000 |
| Lead | 540 | 15 | 140 | 160 | 50 | 420 | 370 |
| Mercury | 12 | 0.2 | 0.5 | 0.43 | 0.5 | 0.55 | 1.6 |
| Nickel | 8200 | 330 | 220 | 130 | 1000 | 15000 | 330 |
| Selenium | 5000 | 450 | 760 | 600 | 1000 | 630 | 720 |
| Silver | 320 | 29 | 220 | --- | 100 | 5500 | --- |
| Thallium | 1700 | 7.9 | 27 | 20 | 20 | 87 | 50 |
| Zinc | 32000 | 140 | 1700 | 2100 | 500 | 21000 | 26100 |

¹ Mercury determined by EPA Method 1631; all others by EPA Method 1638² Lowest ambient water quality criterion (WQC) in the National Toxics Rule (40 CFR 131.36)

Table 8. Comparison of Single-laboratory IDEs resulting from all model types for EPA Methods 1631 and 1638

| Analyte | IDE, Based on Given Model | | | |
|----------|---------------------------|------------------|-------------|------------------------|
| | Constant | Linear | Exponential | Hybrid |
| Antimony | 2400 | -72 ¹ | 140 | 99 |
| Cadmium | 1200 | 120 | 150 | 140 |
| Copper | 2600 | 960 | 780 | 690 |
| Lead | 380 | 140 | 140 | 140 |
| Mercury | 8.1 | 0.059 | 0.79 | 0.50 |
| Nickel | 6600 | -37 ¹ | 220 | 110 |
| Selenium | 4300 | 680 | 760 | 500 |
| Silver | 2400 | 670 | 220 | undefined ² |
| Thallium | 220 | 21 | 27 | 16 |
| Zinc | 9600 | 1500 | 1700 | 1700 |

¹ Negative due to negative intercept estimate in precision model.² IDE did not converge to a single value for estimated models.

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|---------------------------|--------|-----------|-------------|------------|--------------------------|--------------------|-----------------------|
| 1,1,1,2-tetrachloroethane | 502.2 | ELCD | 0.028 | 0.011 | 2.616 | Exponential | Linear |
| 1,1,1,2-tetrachloroethane | 524.2 | | 0.206 | 0.160 | 1.287 | Exponential | Exponential |
| 1,1,1-trichloroethane | 502.2 | ELCD | 0.035 | 0.035 | 0.974 | Exponential | Exponential |
| 1,1,1-trichloroethane | 524.2 | | 0.268 | 0.033 | 8.033 | Exponential | Hybrid |
| 1,1,2,2-tce+1,2,3-tcp | 502.2 | ELCD | 0.170 | 3.404 | 0.050 | Exponential | Constant |
| 1,1,2,2-tetrachloroethane | 524.2 | | 0.377 | 0.505 | 0.746 | Exponential | Exponential |
| 1,1,2-trichloroethane | 502.2 | ELCD | 0.026 | 0.012 | 2.158 | Exponential | Linear |
| 1,1,2-trichloroethane | 524.2 | | 0.284 | 0.218 | 1.300 | Exponential | Exponential |
| 1,1-dichloroethane | 502.2 | ELCD | 0.066 | 0.022 | 2.971 | Exponential | Exponential |
| 1,1-dichloroethane | 524.2 | | 0.206 | 0.089 | 2.310 | Exponential | Exponential |
| 1,1-dichloroethene | 502.2 | ELCD | 0.193 | 0.068 | 2.834 | Exponential | Exponential |
| 1,1-dichloroethene | 524.2 | | 0.278 | 0.077 | 3.607 | Exponential | Hybrid |
| 1,1-dichloropropanone | 524.2 | | 6.032 | 8.566 | 0.704 | Exponential | Exponential |
| 1,1-dichloropropene | 524.2 | | 0.247 | 0.038 | 6.493 | Exponential | Constant ¹ |
| 1,2,3-trichlorobenzene | 502.2 | ELCD | 0.122 | 0.161 | 0.755 | Exponential | Constant |

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|-----------------------------|--------|-----------|-------------|------------|--------------------------|--------------------|-----------------------|
| 1,2,3-trichlorobenzene | 502.2 | PID | 0.114 | 0.060 | 1.891 | Exponential | Exponential |
| 1,2,3-trichlorobenzene | 524.2 | | 0.259 | 0.139 | 1.854 | Exponential | Exponential |
| 1,2,3-trichloropropane | 524.2 | | 1.206 | 0.230 | 5.241 | Exponential | Constant ¹ |
| 1,2,4-trichlorobenzene | 502.2 | ELCD | 0.077 | 0.096 | 0.798 | Exponential | Constant |
| 1,2,4-trichlorobenzene | 502.2 | PID | 0.124 | 0.070 | 1.774 | Exponential | Exponential |
| 1,2,4-trichlorobenzene | 524.2 | | 0.208 | 0.108 | 1.916 | Exponential | Exponential |
| 1,2,4-trimethylbenzene | 502.2 | PID | 0.123 | 0.137 | 0.897 | Exponential | Constant |
| 1,2,4-trimethylbenzene | 524.2 | | 0.129 | 0.054 | 2.380 | Exponential | Exponential |
| 1,2-dibromo-3-chloropropane | 524.2 | | 1.619 | 0.414 | 3.912 | Exponential | Hybrid |
| 1,2-dibromoethane | 502.2 | ELCD | 0.143 | 0.023 | 6.138 | Exponential | Linear |
| 1,2-dibromoethane | 524.2 | | 0.289 | 0.302 | 0.958 | Exponential | Exponential |
| 1,2-dichlorobenzene | 502.2 | ELCD | 0.053 | 0.053 | 1.007 | Exponential | Linear |
| 1,2-dichlorobenzene | 502.2 | PID | 0.147 | 0.067 | 2.197 | Exponential | Exponential |
| 1,2-dichlorobenzene | 524.2 | | 0.112 | 0.065 | 1.727 | Exponential | Exponential |
| 1,2-dichloroethane | 502.2 | ELCD | 0.037 | 0.018 | 2.054 | Exponential | Exponential |
| 1,2-dichloroethane | 524.2 | | 0.229 | 0.201 | 1.139 | Exponential | Exponential |
| 1,2-dichloropropane | 502.2 | ELCD | 0.037 | 0.083 | 0.440 | Exponential | Constant |
| 1,2-dichloropropane | 524.2 | | 0.221 | 0.206 | 1.072 | Exponential | Exponential |
| 1,3,5-tmb+4-chlorotoluene | 502.2 | PID | 0.108 | 0.135 | 0.797 | Exponential | Constant |
| 1,3,5-trimethylbenzene | 524.2 | | 0.117 | 0.045 | 2.612 | Exponential | Exponential |
| 1,3-dichlorobenzene | 502.2 | ELCD | 0.100 | 0.590 | 0.169 | Exponential | Constant |
| 1,3-dichlorobenzene | 502.2 | PID | 0.123 | 0.189 | 0.654 | Exponential | Constant |
| 1,3-dichlorobenzene | 524.2 | | 0.126 | 0.040 | 3.123 | Exponential | Exponential |
| 1,3-dichloropropane | 502.2 | ELCD | 0.045 | 0.218 | 0.204 | Exponential | Constant |
| 1,3-dichloropropane | 524.2 | | 0.170 | 0.112 | 1.517 | Exponential | Exponential |
| 1,4-dichlorobenzene | 502.2 | ELCD | 0.054 | 0.039 | 1.388 | Exponential | Linear |
| 1,4-dichlorobenzene | 524.2 | | 0.110 | 0.047 | 2.333 | Exponential | Exponential |
| 1-chlorobutane | 524.2 | | 0.200 | 0.058 | 3.451 | Exponential | Linear |
| 2,2-dichloropropane | 524.2 | | 0.700 | 0.117 | 5.985 | Exponential | Hybrid |
| 2-butanone | 524.2 | | 0.775 | 1.261 | 0.615 | Exponential | Exponential |
| 2-chlorotoluene | 502.2 | ELCD | 0.184 | 0.104 | 1.776 | Exponential | Exponential |
| 2-chlorotoluene | 502.2 | PID | 0.222 | 0.391 | 0.568 | Exponential | Constant |
| 2-chlorotoluene | 524.2 | | 0.121 | 0.036 | 3.343 | Exponential | Exponential |
| 2-hexanone | 524.2 | | 0.815 | 0.782 | 1.042 | Exponential | Exponential |
| 2-nitropropane | 524.2 | | 0.965 | 8.974 | 0.108 | Exponential | Constant |

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|----------------------|--------|-----------|-------------|------------|--------------------------|--------------------|-------------------|
| 4-chlorotoluene | 502.2 | ELCD | 0.159 | 0.138 | 1.153 | Exponential | Linear |
| 4-chlorotoluene | 524.2 | | 0.102 | 0.039 | 2.648 | Exponential | Exponential |
| 4-isopropyltoluene | 524.2 | | 0.102 | 0.037 | 2.762 | Exponential | Exponential |
| 4-methyl-2-pentanone | 524.2 | | 1.060 | 0.987 | 1.074 | Exponential | Exponential |
| Acetone | 524.2 | | 2.025 | 28.956 | 0.070 | Exponential | Constant |
| Acrylonitrile | 524.2 | | 1.197 | 0.982 | 1.219 | Exponential | Exponential |
| Allyl Chloride | 524.2 | | 0.203 | 0.070 | 2.915 | Exponential | Hybrid |
| Aluminum | 1620 | | 198.565 | 70.438 | 2.819 | Constant | Constant |
| Aluminum | 200.8 | | 12.004 | 21.862 | 0.549 | Exponential | Constant |
| Ammonia as Nitrogen | 350.3 | | 0.013 | 0.038 | 0.336 | Exponential | Constant |
| Antimony | 1620 | | 4.087 | 6.023 | 0.678 | Constant | Linear |
| Antimony | 200.8 | | 0.018 | 0.293 | 0.062 | Exponential | Constant |
| Arsenic | 1620 | | 1.463 | 2.175 | 0.672 | Exponential | Constant |
| Arsenic | 200.8 | | 0.346 | 0.340 | 1.015 | Exponential | Exponential |
| Barium | 1620 | | 1.762 | 1.558 | 1.131 | Constant | Constant |
| Barium | 200.8 | | 0.079 | 0.071 | 1.120 | Exponential | Constant |
| Benzene | 502.2 | PID | 0.077 | 0.058 | 1.324 | Exponential | Exponential |
| Benzene | 524.2 | | 0.115 | 0.030 | 3.907 | Exponential | Exponential |
| Beryllium | 1620 | | 0.428 | 0.427 | 1.003 | Exponential | Exponential |
| Beryllium | 200.8 | | 0.019 | 0.017 | 1.138 | Exponential | Constant |
| Boron | 1620 | | 19.884 | 21.488 | 0.925 | Exponential | Exponential |
| Bromobenzene | 502.2 | ELCD | 0.729 | 0.343 | 2.122 | Linear | Exponential |
| Bromobenzene | 502.2 | PID | 0.048 | 0.025 | 1.947 | Exponential | Exponential |
| Bromobenzene | 524.2 | | 0.175 | 0.156 | 1.125 | Exponential | Exponential |
| Bromochloromethane | 502.2 | ELCD | 0.462 | 0.029 | 15.797 | Linear | Exponential |
| Bromochloromethane | 524.2 | | 0.309 | 0.396 | 0.780 | Exponential | Exponential |
| Bromodichloromethane | 502.2 | ELCD | 0.064 | 0.018 | 3.584 | Exponential | Exponential |
| Bromodichloromethane | 524.2 | | 0.182 | 0.093 | 1.951 | Exponential | Exponential |
| Bromoform | 502.2 | ELCD | 1.450 | 0.023 | 62.020 | Constant | Linear |
| Bromoform | 524.2 | | 0.350 | 0.247 | 1.417 | Exponential | Exponential |
| Bromomethane | 502.2 | ELCD | 6.993 | 0.731 | 9.569 | Constant | Exponential |
| Bromomethane | 524.2 | | 0.238 | 0.148 | 1.610 | Exponential | Linear |
| Cadmium | 1620 | | 0.184 | 0.200 | 0.922 | Exponential | Exponential |
| Cadmium | 200.8 | | 0.011 | 0.015 | 0.735 | Exponential | Constant |
| Calcium | 1620 | | 39.651 | 51.207 | 0.774 | Linear | Constant |

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|-------------------------|--------|-----------|-------------|------------|--------------------------|--------------------|-------------------|
| Carbon Disulfide | 524.2 | | 0.203 | 0.083 | 2.440 | Exponential | Linear |
| Carbon Tetrachloride | 524.2 | | 0.258 | 0.166 | 1.549 | Exponential | Linear |
| Carbontet+1,1-dcp | 502.2 | ELCD | 0.062 | 0.049 | 1.274 | Exponential | Exponential |
| Chloroacetonitrile | 524.2 | | 1.535 | 1.842 | 0.833 | Exponential | Exponential |
| Chlorobenzene | 502.2 | ELCD | 0.440 | 0.044 | 9.900 | Linear | Exponential |
| Chlorobenzene | 502.2 | PID | 0.062 | 0.058 | 1.083 | Exponential | Exponential |
| Chlorobenzene | 524.2 | | 0.119 | 0.035 | 3.378 | Exponential | Exponential |
| Chloroethane | 502.2 | ELCD | 2.492 | 0.090 | 27.544 | Constant | Linear |
| Chloroethane | 524.2 | | 0.336 | 0.239 | 1.403 | Exponential | Exponential |
| Chloroform | 502.2 | ELCD | 0.031 | 0.008 | 3.923 | Exponential | Linear |
| Chloroform | 524.2 | | 0.203 | 0.111 | 1.829 | Exponential | Exponential |
| Chloromethane | 502.2 | ELCD | 0.204 | 0.499 | 0.409 | Exponential | Constant |
| Chloromethane | 524.2 | | 0.215 | 0.148 | 1.456 | Exponential | Exponential |
| Chromium | 1620 | | 0.478 | 0.728 | 0.657 | Exponential | Constant |
| Chromium | 200.8 | | 0.393 | 0.473 | 0.830 | Linear | Constant |
| Cis-1,2-dce+2,2-dcp | 502.2 | ELCD | 0.043 | 0.027 | 1.575 | Exponential | Exponential |
| Cis-1,2-dichloroethene | 524.2 | | 0.203 | 0.181 | 1.123 | Exponential | Exponential |
| Cis-1,3-dichloropropene | 502.2 | ELCD | 0.059 | 0.015 | 3.860 | Exponential | Exponential |
| Cis-1,3-dichloropropene | 502.2 | PID | 0.077 | 0.104 | 0.743 | Exponential | Exponential |
| Cis-1,3-dichloropropene | 524.2 | | 0.158 | 0.103 | 1.536 | Exponential | Exponential |
| Cobalt | 1620 | | 15.560 | 11.825 | 1.316 | Exponential | Exponential |
| Cobalt | 200.8 | | 0.008 | 0.000 | -133.113 | Exponential | Exponential |
| Copper | 1620 | | 20.328 | 15.251 | 1.333 | Constant | Constant |
| Copper | 200.8 | | 0.770 | 0.873 | 0.882 | Constant | Constant |
| Dibromochloromethane | 502.2 | ELCD | 0.418 | 0.380 | 1.098 | Linear | Constant |
| Dibromochloromethane | 524.2 | | 0.253 | 0.192 | 1.315 | Exponential | Exponential |
| Dibromomethane | 502.2 | ELCD | 0.441 | 0.286 | 1.543 | Linear | Constant |
| Dibromomethane | 524.2 | | 0.342 | 0.347 | 0.985 | Exponential | Exponential |
| Dichlorodifluoromethane | 502.2 | ELCD | 0.087 | 1.175 | 0.074 | Exponential | Constant |
| Dichlorodifluoromethane | 524.2 | | 0.420 | 0.265 | 1.584 | Exponential | Exponential |
| Diethyl Ether | 524.2 | | 0.340 | 0.314 | 1.082 | Exponential | Exponential |
| Ethyl Methacrylate | 524.2 | | 0.244 | 0.243 | 1.003 | Exponential | Exponential |
| Ethylbenzene | 502.2 | PID | 0.075 | 0.048 | 1.571 | Exponential | Exponential |
| Ethylbenzene | 524.2 | | 0.167 | 0.111 | 1.504 | Exponential | Exponential |
| Hardness | 130.2 | | 2.152 | 4.687 | 0.459 | Exponential | Constant |

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|--------------------------------|---------------|------------------|--------------------|-------------------|----------------------------------|----------------------------|---------------------------|
| Hexachlorobutadiene | 502.2 | ELCD | 0.090 | 0.069 | 1.300 | Exponential | Linear |
| Hexachlorobutadiene | 524.2 | | 0.263 | 0.208 | 1.265 | Exponential | Exponential |
| Hexachloroethane | 524.2 | | 0.234 | 0.234 | 1.000 | Exponential | Exponential |
| Hexchlorobutadiene+naphthalene | 502.2 | PID | 0.598 | 0.567 | 1.055 | Exponential | Constant |
| Iron | 1620 | | 345.686 | 1021.716 | 0.338 | Linear | Constant |
| Isopropylbenzene | 502.2 | PID | 0.059 | 0.038 | 1.551 | Exponential | Exponential |
| Isopropylbenzene | 524.2 | | 0.113 | 0.036 | 3.137 | Exponential | Exponential |
| Lead | 1620 | | 2.317 | 2.831 | 0.819 | Exponential | Constant |
| Lead | 200.8 | | 0.197 | 2.771 | 0.071 | Exponential | Constant |
| M+p Xylene | 502.2 | PID | 0.116 | 0.114 | 1.014 | Exponential | Constant |
| M+p Xylene | 524.2 | | 0.127 | 0.033 | 3.846 | Exponential | Exponential |
| Magnesium | 1620 | | 99.662 | 176.736 | 0.564 | Exponential | Constant |
| Manganese | 1620 | | 6.531 | 4.363 | 1.497 | Constant | Constant |
| Manganese | 200.8 | | 0.106 | 0.074 | 1.419 | Constant | Constant |
| Mercury | 200.8 | | 0.062 | 0.014 | 4.550 | Exponential | Hybrid |
| Methacrylonitrile | 524.2 | | 0.643 | 0.529 | 1.216 | Exponential | Hybrid |
| Methyl Iodide | 524.2 | | 0.173 | 0.099 | 1.746 | Exponential | Exponential |
| Methyl Tert-butyl Ether | 524.2 | | 0.195 | 0.154 | 1.266 | Exponential | Exponential |
| Methylacrylate | 524.2 | | 0.549 | 0.551 | 0.995 | Exponential | Exponential |
| Methylene Chloride | 502.2 | ELCD | 2.727 | -1.325 | -2.057 | Constant | Constant |
| Methylene Chloride | 524.2 | | 0.276 | 0.159 | 1.737 | Exponential | Exponential |
| Methylmethacrylate | 524.2 | | 0.484 | 0.366 | 1.322 | Exponential | Linear |
| Molybdenum | 1620 | | 2.917 | 5.783 | 0.504 | Exponential | Constant |
| Molybdenum | 200.8 | | 0.262 | 0.005 | 47.826 | Constant | Constant |
| N-butylbenzene | 502.2 | PID | 0.139 | 0.049 | 2.819 | Exponential | Linear |
| N-butylbenzene | 524.2 | | 0.136 | 0.052 | 2.605 | Exponential | Exponential |
| N-propylbenzene | 502.2 | PID | 0.089 | 0.100 | 0.889 | Exponential | Constant |
| N-propylbenzene | 524.2 | | 0.231 | 0.050 | 4.596 | Exponential | Hybrid |
| Naphthalene | 524.2 | | 0.175 | 0.180 | 0.968 | Exponential | Exponential |
| Nickel | 1620 | | 23.784 | 40.205 | 0.592 | Exponential | Constant |
| Nickel | 200.8 | | 0.076 | 0.068 | 1.124 | Exponential | Constant |
| o-xylene | 524.2 | | 0.161 | 0.084 | 1.925 | Exponential | Exponential |
| o-xylene+styrene | 502.2 | PID | 0.111 | 0.145 | 0.769 | Exponential | Constant |
| P-isoproptol+1,4-dcb | 502.2 | PID | 0.153 | 0.143 | 1.068 | Exponential | Constant |
| Pentachloroethane | 524.2 | | 0.337 | 0.419 | 0.805 | Exponential | Linear |

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|-----------------------------|--------|-----------|-------------|------------|--------------------------|--------------------|-----------------------|
| Sec-butylbenzene | 502.2 | PID | 0.079 | 0.053 | 1.504 | Exponential | Exponential |
| Sec-butylbenzene | 524.2 | | 0.132 | 0.038 | 3.422 | Exponential | Exponential |
| Selenium | 1620 | | 1.915 | 1.757 | 1.090 | Exponential | Exponential |
| Selenium | 200.8 | | 0.410 | 0.311 | 1.317 | Exponential | Exponential |
| Silver | 1620 | | 10.219 | 11.118 | 0.919 | Exponential | Constant |
| Silver | 200.8 | | 0.010 | -0.030 | -0.348 | Exponential | Exponential |
| Sodium | 1620 | | 133.007 | 134.645 | 0.988 | Exponential | Exponential |
| Styrene | 524.2 | | 0.119 | 0.045 | 2.676 | Exponential | Exponential |
| Tert-butylbenzene | 502.2 | PID | 0.073 | 0.049 | 1.498 | Exponential | Exponential |
| Tert-butylbenzene | 524.2 | | 0.170 | 0.055 | 3.093 | Exponential | Exponential |
| Tetrachloroethene | 502.2 | ELCD | 0.051 | 0.045 | 1.119 | Exponential | Exponential |
| Tetrachloroethene | 502.2 | PID | 0.157 | 0.097 | 1.607 | Exponential | Linear |
| Tetrachloroethene | 524.2 | | 0.379 | 0.526 | 0.721 | Exponential | Linear |
| Thallium | 1620 | | 1.208 | 1.198 | 1.009 | Exponential | Linear |
| Thallium | 200.8 | | 0.001 | 0.000 | 3.140 | Exponential | Exponential |
| Thorium | 200.8 | | 0.003 | 0.000 | 7.428 | Exponential | Constant |
| Tin | 1620 | | 1.200 | 4.383 | 0.274 | Exponential | Exponential |
| Titanium | 1620 | | 5.238 | 19.982 | 0.262 | Exponential | Constant |
| Toluene | 502.2 | PID | 0.061 | 0.061 | 0.999 | Exponential | Constant |
| Toluene | 524.2 | | 0.130 | 0.142 | 0.918 | Exponential | Constant ¹ |
| Total Phosphorus | 365.2 | | 0.013 | 0.011 | 1.219 | Exponential | Exponential |
| Total Suspended Solids | 160.2 | | 2.877 | 2.131 | 1.350 | Exponential | Exponential |
| Trans-1,2-dichloroethene | 502.2 | ELCD | 0.065 | 0.061 | 1.069 | Exponential | Linear |
| Trans-1,2-dichloroethene | 524.2 | | 0.255 | 0.072 | 3.531 | Exponential | Hybrid |
| Trans-1,3-dichloropropene | 502.2 | ELCD | 0.082 | 0.021 | 3.845 | Exponential | Exponential |
| Trans-1,3-dichloropropene | 502.2 | PID | 0.085 | 0.110 | 0.774 | Exponential | Exponential |
| Trans-1,3-dichloropropene | 524.2 | | 0.204 | 0.127 | 1.613 | Exponential | Exponential |
| Trans-1,4-dichloro-2-butene | 524.2 | | 1.182 | 1.205 | 0.981 | Exponential | Exponential |
| Trichloroethene | 502.2 | ELCD | 0.050 | 0.013 | 3.715 | Exponential | Exponential |
| Trichloroethene | 502.2 | PID | 0.096 | 0.080 | 1.189 | Exponential | Exponential |
| Trichloroethene | 524.2 | | 0.288 | 0.329 | 0.875 | Exponential | Linear |
| Trichlorofluoromethane | 502.2 | ELCD | 1.997 | 0.660 | 3.025 | Constant | Constant |
| Trichlorofluoromethane | 524.2 | | 0.307 | 0.290 | 1.059 | Exponential | Exponential |
| Uranium | 200.8 | | 0.001 | 0.000 | 4.743 | Exponential | Exponential |
| Vanadium | 1620 | | 10.063 | 8.643 | 1.164 | Exponential | Exponential |

Table 9. Comparison of 16-point and 5-point Single-laboratory IDEs (SL-IDEs) for the Episode 6000 Data Set

| Analyte | Method | Procedure | SL-IDE (16) | SL-IDE (5) | SL-IDE(16)/ SL-IDE(5) | SL-IDE 16 Model | SL-IDE 5 Model |
|----------------|--------|-----------|-------------|------------|--------------------------|--------------------|-------------------|
| Vanadium | 200.8 | | 0.845 | 0.987 | 0.856 | Exponential | Linear |
| Vinyl Chloride | 502.2 | ELCD | 3.521 | 0.369 | 9.543 | Constant | Linear |
| Vinyl Chloride | 524.2 | | 0.295 | 0.180 | 1.641 | Exponential | Linear |
| WAD Cyanide | 1677 | | 0.672 | 1.243 | 0.541 | Linear | Constant |
| Xylene (Total) | 524.2 | | 0.111 | 0.030 | 3.735 | Exponential | Exponential |
| Yttrium | 1620 | | 3.119 | 13.404 | 0.233 | Exponential | Constant |
| Zinc | 1620 | | 4.415 | 6.661 | 0.663 | Exponential | Constant |
| Zinc | 200.8 | | 1.497 | 5.061 | 0.296 | Exponential | Constant |

Note: ELCD or PID in the Procedure column indicates the photo-ionization detector (PID) or electrolytic conductivity detector (ELCD) in EPA Method 502.2

¹ Original model picked was Hybrid, but failed to converge

Summary Statistics for Table 9

| | IDE(16)/IDE(5) (all analytes) | IDE(16)/IDE(5) (same model used) | IDE(16)/IDE(5) (different models used) |
|--|----------------------------------|-------------------------------------|---|
| Number of Analytes | 198 | 108 | 90 |
| Minimum: | -133.113 | -133.113 | 0.050 |
| 25th percentile: | 0.876 | 1.072 | 0.564 |
| Median: | 1.270 | 1.419 | 0.919 |
| 75th percentile: | 2.328 | 2.310 | 2.122 |
| Maximum: | 62.020 | 47.826 | 62.020 |
| | | | |
| | Number of analytes | Percent of analytes | p-value (for percent=50) |
| SL-IDE (16) > SL-IDE (5) (all analytes) | 133 | 67.17% | <0.0001 |
| SL-IDE (16) > SL-IDE (5) (same model used) | 91 | 84.26% | <00001 |
| SL-IDE (16) > SL-IDE (5) (different models used) | 43 | 48.31% | 0.832 |